

# Agricultural trade: Impacts on food security, groundwater and energy use

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## Addresses

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## Introduction

Trade in food and agricultural products has more than doubled in real terms since 1995. Emerging and developing economies have joined countries in the Global North as dynamic participants in global markets, and now account for about a third of global trade [1]. In general, countries in Latin America, East Africa, and South Asia are net food exporters while most of the rest in Asia and Africa have remained food importers (Figure 1).

Global production of primary crops increased by 53% between 2000 and 2019, to a record high of 9.4 billion tonnes in 2019. This increase in production has been mostly due to intensified use of irrigation, pesticides, and fertilisers, and to a lesser extent to larger cultivated areas, better farming practices, and high-yield crops [2].

Increasing food exports have strengthened local and global food security. On the negative side, they have severely harmed freshwater ecosystems. Surface and groundwater exploitation have resulted in saline intrusion, declining groundwater tables that affect other users, land subsidence, reduction of flow in streams, lakes and wetlands, loss of biodiversity, and heightened

greenhouse gas emissions. Intensive use of pesticides and fertilisers, as well as chemicals, pharmaceuticals and pathogens introduced due to untreated sewage, have also caused serious pollution [1,4].

Globally, most of the water that is used (60%–70%) is groundwater. Its use in irrigation has increased both as a percentage and in absolute terms. It was calculated at 820 km<sup>3</sup>/year in 2018 based on aggregated country-level reports the same year [5]. It is estimated that of all the area equipped for irrigation, over 30% depends on groundwater [6].

Studies reveal large impacts of global food demands on local freshwater resources. As agriculture becomes more water-intensive, more water is embedded in its produce, with growing international imports and exports representing a growing ‘trade’ in this ‘virtual water’ [7–11]. While this contributes indirectly towards food security, it directly affects quantity and quality of water in food-producing arid and semi-arid regions [12,13]. For these regions, it would be more sustainable, and profitable, to import agricultural products and the associated virtual water, from regions that are not water scarce. This would reduce, and even avoid, groundwater depletion and pollution in countries of destination but not in those of origin.

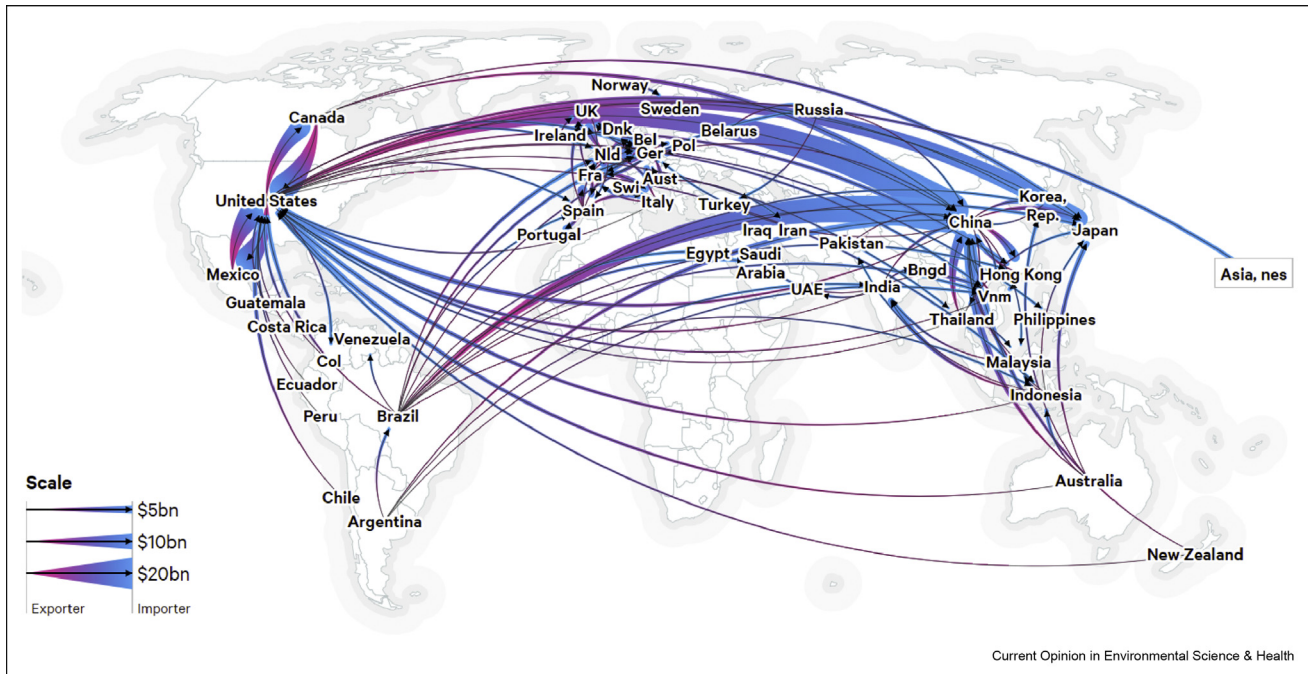
Groundwater’s invisibility has led decision-makers to neglect its management. This includes conjunctive use of surface and groundwater for more efficient management, and coordination of policies among different sectors, such as water, irrigation, and energy. It also causes the various sectors to make decisions that disadvantage other sectors, often for political reasons.

In this analysis, we consider the impacts of agricultural trade on food security, groundwater, and energy use. We report on the current state of groundwater depletion in specific countries followed by an analysis on energy use, and broader policies that have the objective to protect groundwater in more countries around the world.

## Groundwater-based irrigated agriculture for export

There is a clear global trend towards expansion of irrigated areas for export agriculture, which increases productivity but also results in groundwater depletion. Poor management of groundwater in countries in both the

Figure 1



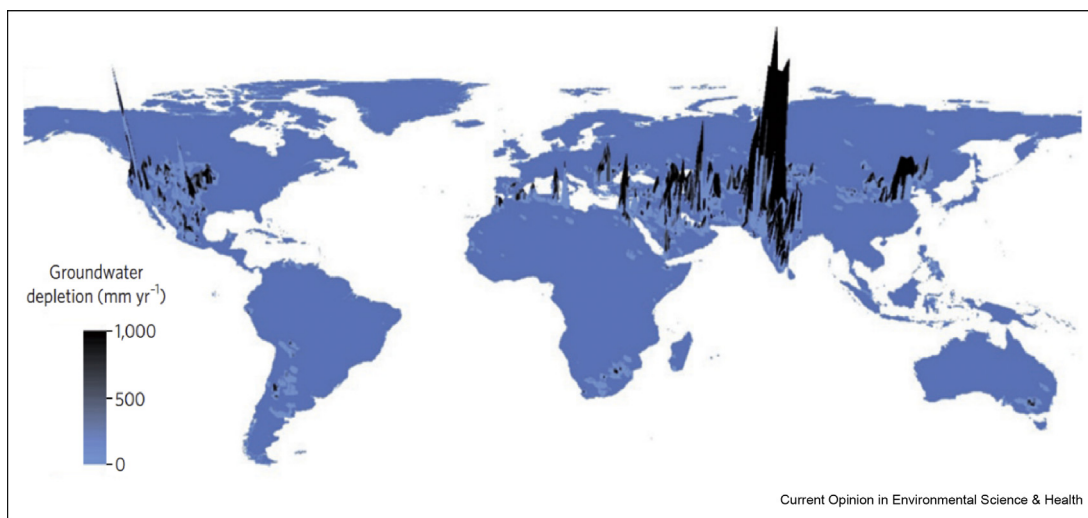
Export of agricultural products in 2015 by value (USD), map by Applied Works [3].

Global South and North has resulted in localised depletion of aquifers (Figure 2), seasonal exhaustion, and pollution.

A country where agricultural production depends greatly on aquifers that are already depleted and polluted is India [16]. It is the largest user of groundwater for irrigation and a main contributor to the global food

basket. Approximately 90% of the groundwater that is extracted is used to irrigate 60% of land area through more than 21 million privately owned wells [17], supporting more than 90 million rural households [18]. India produces approximately 10% of all global agricultural outputs and is the second-largest producer of wheat and rice. From April 2020 to February 2021, the country exported pulses and dairy products worth US

Figure 2



Recent estimate of the global distribution of groundwater depletion. The three-dimensional topography [14] shows ‘mountains of groundwater depletion’ especially in the United States, Mexico, Saudi Arabia, Pakistan, India, and China [15].

\$261 million and US \$183 million, respectively. The total value of its agricultural exports is on track to grow to US \$60 billion by 2022. Nevertheless, most of these exports are grown in the arid or semi-arid parts of the country [19] and rely on groundwater that is over-exploited and polluted. It is estimated that, in 2010, India exported approximately 25 km<sup>3</sup> of virtual water in its agricultural exports. This is equivalent to annual water demand of approximately 13 million people [20], threatening water and food security and local livelihoods [21].

In the United States, also one of the most important countries in global food trade, despite recent gains in water use efficiency, groundwater withdrawal for irrigation has almost tripled since records began in 1950. Consequently, depleting groundwater in many more areas across the country [22]. The High Plains, known as the 'grain basket', and the Central Valley, known as the 'fruit and vegetable basket', are ranked first and second, respectively, among aquifers in the United States for total groundwater withdrawals [23]. In 2019, California exported approximately 28% (by volume) of its agricultural production, earning \$21.71 billion in sales [24]. Its dependence on groundwater for irrigation continues to increase, threatening further aquifer sustainability, and therefore future crop production.

In Mexico, over the past three decades, the state of Guanajuato, in the highlands of Central Mexico, has become the leading supplier of fresh vegetables and fruits to the United States [25]. In 2016, irrigation of 250,000 ha consumed 84% of all extracted groundwater [26]. In 2017, more than 4000 Mm<sup>3</sup> of water was used for irrigation. Unsustainable practices have resulted in overdraft of more than 1000 Mm<sup>3</sup>/year, with pumping depths sinking to more than 200 m below the surface, and aquifer levels dropping on average 2–3 m/year [27]. Growing arsenic and fluoride contamination threatens public health, and, as in California and India, land subsidence that affects channels and drains are common [28]. As in many other regions, agricultural production and intensive irrigation have not considered the recharge capacity of aquifers, which are now severely depleted [29]. If the same practices continue, the estimated balances for 2036 show very large deficits, putting food production at risk [30].

In Ica Province in Peru, the agro-export companies have transformed the desert into farmland. Crop production for export (especially asparagus) depends on groundwater extraction from the Ica-Villacurí aquifer. The production area increased from 7400 ha in 1997 to 22,000 in 2013. As a result, groundwater use more than tripled within this period [31], and the groundwater table has sunk by 1.5–4 m per year [32]. In 2012, surface water demand totalled 250 Mm<sup>3</sup> for more than 5800 users, while demand for groundwater was 300 Mm<sup>3</sup> for 10 users, all of

them companies, threatening small farmers and domestic water users [33]. With the objective to protect the aquifer, well digging has been prohibited. However, in the absence of monitoring, construction of illegal wells has increased. It has been estimated that 76% of the cultivated area in the Ica Valley will suffer severe water shortages within 10 years.

In Spain, expansion of groundwater-based irrigation (3.8 million ha in 2020) has brought significant socioeconomic development [34] but also increased deterioration and pollution of groundwater in specific regions and basins. In 2019, the country exported 14 million tonnes of fruit and vegetables (7.7% more than in 2018), mainly to elsewhere in Europe, earning over €13 billion (5.5% more than in 2018). Irrigation systems have become more efficient, but they have not resulted in water savings. This is because irrigated area has expanded with the proportional use of groundwater for irrigation growing from 17.5% to 24%, and the volume rising from 3189 Mm<sup>3</sup>/year to 4142 Mm<sup>3</sup>/year [35] (30% increase in water abstraction), exacerbating overdrafts.

Water efficiency alone has not been the solution. There is a rebound effect that encourages greater use of resources: more land, more water, more fertilisers [36]. Examples include China [37,38], Spain [39], Australia [40], India [41] and so on. Without incentives to moderate water consumption, there is a strong risk of over-exploitation and even depletion of water resources. The importance of groundwater is recognised. However, policies to remediate environmental problems have been slow to be implemented nearly all over the world [42].

### Groundwater extraction and energy use

Groundwater resources typically use about 30% more energy than do surface water supplies [43]. Subsidies on fuel (diesel, petrol, butane) or grid electricity for agricultural use have resulted in over-abstraction of surface and ground waters. They have driven inefficient and energy-intensive water use by hiding the true cost of the resources [44]. They have also been very expensive for governments. In Morocco, they have represented up to 6% of the national GDP [45].

India is the fourth-largest energy consumer globally, partly due irrigation pumping, encouraged by subsidies. In 2003–2004, around 12.8 million electric pumps, with a total of 52 GW (GW) of connected load, consumed 87 billion kilowatt-hours (kWh) of electricity. In 2017, irrigation consumed 17% of energy produced at the national level [46]. Energy subsidies for agricultural groundwater pumping represent up to 85% of the actual cost of electricity. Without specific objectives and time frames, subsidies threaten the sustainability of both groundwater and the power sector.

In the North China Plain, where there are severe issues with aquifer depletion, groundwater pumping consumes an average of 13.67 billion kWh/year, and 1122 kWh/Mm<sup>3</sup> under the winter wheat/summer maize rotation system. The region has become one of the world's largest energy consumers for groundwater irrigation [47].

In the US, in California, it is estimated that 7000 GWh of electricity were used for groundwater extraction in 2010. More specifically, the Santa Clara Valley Water District estimates that farmers in the San Francisco Bay Area used about 1000 kWh for 4546 m<sup>3</sup> (1000 kWh/million gallons) for groundwater pumping [48].

While groundwater has historically been an inexpensive resource for farmers, especially where groundwater use is not regulated, rising energy prices have become a substantial problem. Even though farmers can spend as much as 25% of their average annual net income from crops on more powerful pumps to pump water out of deeper wells [49], there are no indications that rising prices are slowing groundwater withdrawals [50].

As an alternative to conventional electric and diesel pumping systems, several countries are promoting subsidised solar-powered irrigation. Their wider use will allow small and marginal farmers to pump their own groundwater instead of buying expensive water from large farmers. Solar power can also reduce GHG emissions from agriculture. On the other hand, without regulation, solar pumps will contribute to the depletion of groundwater (even if irrigation efficiency is maximised), making it unsustainable. Farmers using solar power have no financial incentive to limit their water pumping [51,52]. They can reallocate water to larger areas of land, more water-intensive crops, an additional cropping season, or higher yields. Some may sell their 'extra' water to neighbours, putting more pressure on already scarce water resources.

Given the impacts of groundwater extraction on energy, energy embedded in water savings could be measured to inform policies related to water and energy efficiency [53]. Decision makers could also consider the energy implications of water policy decisions, improve coordination among resource management agencies, assign a higher priority to water conservation and work closer with the farmers [37,54].

### **The future of food security under unsustainable groundwater management**

Groundwater management for irrigated agriculture depends on policy, legal and regulatory frameworks, and on their implementation. It also requires an enabling environment—that is, institutional capacity and collaboration in public institutions in the various sectors (water, agriculture, energy, environment) and at all levels: federal offices and state boards, basin

authorities, water users' and farmers' organisations, as well as private sector groups.

Reasons for the current situation of groundwater have been discussed before. They include poor regulation, policy, management, and governance; institutions without capacity or resources (human as well as financial) to implement plans and policies; lack of realistic and informed goals that consider aquifer roles and uses as well as limited or unreliable data and information; absence of effective processes to engage users; and the fact that politicians prefer not to increase water prices or reduce subsidies because of possible repercussions in terms of their electability.

In the Global South, long-term action plans to protect groundwater resources that are implementable, are mostly lacking [42]. There are regulations that either limit or prohibit the use of specific aquifers, but the results have been limited [55,56]. China's first national plan on groundwater pollution control [57] and India's Atal Bhujal Yojana—National Groundwater Management Improvement Programme [58] are examples of recent initiatives. If implemented, they will result in more effective groundwater protection.

In the Global North, there are more efforts to protect groundwaters. In the United States, in California, the Sustainable Groundwater Management Act was passed in 2014, with full implementation scheduled for 2041 [59]. The Act aims to stop overdraft and bring groundwater basins into balanced levels of pumping and recharge. No such plans exist in Texas, one of the five states that use the most groundwater in the country, where groundwater is not regulated.

In the European Union, the Groundwater Directive considers several measures to achieve good quantitative and chemical status of groundwater by 2015 [60]. Regulations are at different levels of implementation in the member states. The EU has also approved a regulation on minimum requirements for water reuse for agricultural irrigation [61]. Reused water is proposed as an alternative resource to improve the status of the environment and alleviate pressure on groundwater by substituting abstraction as well as by relieving pressure of discharge to sensitive areas.

In Australia, groundwater is managed by state and territory governments with the help of regulatory and economic instruments [62]. However, political decisions have resulted in poor outcomes, particularly regarding investment capital [63].

Policy interventions to protect groundwater are essential. However, in many places, overexploitation has resulted in a growing gap between water extraction and recharge, loss of surface water resources as part of the

same ecological system, reduction of water available for the environment, rapid decay of traditional irrigation sources such as tanks and spring channels, stagnant irrigation canals, progressive decline of groundwater tables and deterioration of water quality. Policies to reduce groundwater extraction include water accounting, pricing water considering all its uses (including environmental and opportunity costs), higher electricity prices, water use rationing (with quantitative ceilings on water and electricity use/ha), promoting recharge, limit drilling depth and proximity of wells, encouraging farmers to grow less water-intensive crops, information campaigns, use of social media, and so on. However, without laws and regulations that are implementable, and more effective groundwater management that assesses and limits exploitation, limits irrigation area, controls illegal irrigation and use of fertilisers and pesticides, and monitors water quantity and quality, pollution and rapid depletion of aquifers have become the norm in many parts of the world.

The potential of digitalisation for a more effective management of groundwater quantity and quality is enormous. However, benefits are uneven across developing countries as availability and access to data and infrastructure are limited.

Despite depletion and pollution, use of groundwater for irrigation has benefitted local and national economies. Thus, for governments in OECD and non-OECD countries it would be almost impossible to withdraw agricultural subsidies, estimated at \$700 billion/year, \$536 billion of them directly to producers [64]. This, even when eliminating minimum prices or subsidies in arid or semi-arid countries for crops that are water intensive would greatly contribute to a better management of water resources [42].

Overall, we see a world with a growing gap between demand for and availability of water at local levels, with intensive agricultural practices contributing to the severe degradation of the environment, and growing virtual water trade, which in turn can threaten both livelihoods and freshwater ecosystems in origin. Additionally, the increasing frequency and duration of droughts makes it essential for public and private institutions and users alike, to understand and improve management practices. Because groundwater is crucial for agriculture, over-exploitation, coupled with drought events, has increased vulnerability at the global level.

Scholars suggest that, to adapt to climate change, regions and countries should grow less water-intensive crops [65], or convert irrigated agriculture to rainfed agriculture, even for valuable crops [66]. Neither of these changes is likely to be attractive to farmers, compared to the reliability that irrigation provides.

Urgently needed are groundwater governance mechanisms with joint management of resources and appropriate forms of shared decision-making by actors at the national and local levels, and goals that are broadly agreed and locally implemented—features lacking at present.

As aquifers continue to be depleted, and global climate continues to change, pumping groundwater could become economically prohibitive and environmentally more damaging, further affecting food security, freshwater ecosystems, and livelihoods. Solutions are in the hands of governments and users. Inaction will have serious consequences for agricultural sustainability, long-term food security, community livelihoods, and economic growth.

### Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

### References

Papers of particular interest, published within the period of review, have been highlighted as:

\* of special interest

1. *The state of agricultural commodity markets: agricultural markets and sustainable development: global value chains, smallholder farmers and digital innovations*. Rome: Food and Agriculture Organization; 2020. Available at, <http://www.fao.org/documents/card/en/cb0665en/>.
2. *Statistical yearbook. World food and agriculture 2021*. Rome: Food and Agriculture Organization; 2021. Available at, <https://reliefweb.int/sites/reliefweb.int/files/resources/cb4477en.pdf>.
3. Benton T: *Food security, trade and its impacts*. 2017. Available at, <https://resourcetrade.earth/publications/food-security-trade-and-its-impacts>.
4. Lall U, Josset L, Russo T: **A snapshot of the world's groundwater challenges**. *Annu Rev Environ Resour* 2020, **45**:171–194. Available at, <https://www.annualreviews.org/doi/pdf/10.1146/annurev-environ-102017-025800>.
5. *The state of the world's land and water resources for food and agriculture – systems at breaking point. Synthesis report 2021*. Rome: Food and Agriculture Organization; 2021. Available at, <https://www.fao.org/3/cb7654en/cb7654en.pdf>.
6. Siebert S, Burke J, Faures MJ, Frenken K, Hoogeveen J, Döll P, Portmann FT: **Groundwater use for irrigation—a global inventory**. *Hydrol Earth Syst Sci* 2010, **14**:1863–1880, <https://doi.org/10.5194/hess-14-1863-2010>.  
This is the most recent global inventory of irrigated areas globally and the related consumptive water uses are determined.
7. Dalin C, Wada Y, Kastner T, Puma MJ: **Groundwater depletion embedded in international food trade**. *Nature* 2017, **543**:700–704, <https://doi.org/10.1038/nature21403>.
8. Graham NT, Hejazi MI, Kim SH, Davies EGR, Edmonds JA, Miralles-Wilhelm F: **Future changes in the trading of virtual water**. *Nat Commun* 2020, **11**:3632, <https://doi.org/10.1038/s41467-020-17400-4>.
9. Marston L, Konar M: **Drought impacts to water footprints and virtual water transfers of the Central Valley of California**. *Environ Res Lett* 2017, **53**:5756–5773, <https://doi.org/10.1002/2016WR020251>.

10. Marston L, Konar M, Cai X, Troy TJ: **Virtual groundwater transfers from overexploited aquifers in the United States.** *Proc Natl Acad Sci Unit States Am* 2015, **112**:8561–8566. Available at, <https://www.pnas.org/content/112/28/8561>.
- A very comprehensive analysis that traces virtual groundwater transfers.
11. Rosa L, Chriarelli DD, Tu C, Rulli MC, ODorico P: **Global unsustainable virtual water flows in agricultural tards.** *Environ Res Lett* 2019, **14**:114001.
12. Aldaya MM, Allan JA, Hoekstra AY: **Strategic importance of green water in international crop trade.** *Ecol Econ* 2010, **69**:887–894.
13. Chico D, Aldaya MM, Flachsbarth I, Garrido A: **Virtual water trade, food security and sustainability: lessons from Latin America and Spain.** In *Integrated water resources management in the 21st century: revisiting the paradigm*. Edited by Martínez-Santos P, Aldaya MM, Llamas MR, CRC Press/Balkema; 2014:75–98.
14. Wada Y, van Beek LPH, Weiland FCS, Chao BF, Wu YH, Bierkens MFP: **Past and future contribution of global groundwater depletion to sea-level rise.** *Geophys Res Lett* 2012, **39**:L09402. Available at, <https://agupubs.onlinelibrary.wiley.com/doi/full/10.1029/2012GL051230>.
15. Werner AH, Gleeson T: **Regional strategies for the accelerating global problem of groundwater depletion.** *Nature* 2012, **5**:853–861. Available at, <https://www.nature.com/articles/ngeo1617>.
16. Kumar MD, Saleth RM, Foster D, Niranjana V, Sivamohan MVK: **Water, human development, inclusive growth, and poverty alleviation: international perspectives.** In *Rural water systems for multiple uses and livelihood security*. Edited by Kumar MD, James AJ, Kabir Y, Elsevier Science; 2016:17–47.
17. Gulati A, Sharma B, Banerjee P, Mohan G: *Getting more from Less: story of India's shrinking water resources.* Delhi: Indian Council for Research on International Economic Relations; 2019.
18. *National compilation on dynamic ground water resources of India. Faridabad: central ground water board.* Department of Water Resources, River Development & Ganga Rejuvenation. Ministry of Jal Shakti, Government of India; 2021.
19. Amarasinghe UA, Bharat RS, Aloysius N, Scott C, Smakhtin V, De Fraiture C, Sinha AK, Shukla AK: *Spatial variation in water supply and demand across river basins of India research [Report 83].* Sri Lanka: International Water Management Institute; 2004.
20. Dhawan V: *Water and agriculture in India. Background paper for the South Asia expert panel during the global forum for food and agriculture.* Hamburg: German Asia-Pacific Business Association; 2017. Available at, [https://www.oav.de/fileadmin/user\\_upload/5\\_Publikationen/5\\_Studien/170118\\_Study\\_Water\\_Agriculture\\_India.pdf](https://www.oav.de/fileadmin/user_upload/5_Publikationen/5_Studien/170118_Study_Water_Agriculture_India.pdf).
21. Jain M, Fishman R, Mondal P, Galford GL, Bhattarai N, Naeem S, Lall U: **Balwinder-Singh, DeFries RS: groundwater depletion will reduce cropping intensity in India.** *Sci Adv* 2021, **7**:1–9, <https://doi.org/10.1126/sciadv.abd2849>.
- It is a very good analysis of groundwater depletion in India
22. Scanlon BR, Faunt CC, Longuevergne L, Reedy RC, Alley WM, McGuire VL, McMahon PB: **Groundwater depletion and sustainability of irrigation in the US high Plains and Central Valley.** *Proc Natl Acad Sci Unit States Am* 2012, **109**:9320–9325, <https://doi.org/10.1073/pnas.1200311109>.
23. Maupin MA, Barber NL: *Estimated withdrawals from principal aquifers in the United States, 2000.* Reston, VA: US Geological Survey; 2005.
24. Tortajada C, Kastner MJ, Buurman J, Biswas AK: **The California drought: coping responses and resilience building.** *Environ Sci Pol* 2017, **78**:97–113, <https://doi.org/10.1016/j.envsci.2017.09.012>.
25. Evaluación de impactos de la tecnificación del riego en Guanajuato. Guanajuato: secretaría de Desarrollo Agroalimentario y Rural; n.d. Available at, <https://sdayr.guanajuato.gob.mx/contenido/adjuntos/evaluaciones/2020/EvaluaciondelaTecnificaciondelRiegoenElEstadodeGuanajuato.pdf>.
26. Hoogesteger J, Massink G: **Corporate labour standards and work quality: insights from the agro-export sector of Guanajuato, Central Mexico.** *Third World Q* 2021, **42**: 1196–1212, <https://doi.org/10.1080/01436597.2021.1874334>.
27. *Compendio de Agua Subterránea en Guanajuato. Guanajuato: comisión Estatal del Agua.* 2018.
28. Hoogesteger J, Wester P: **Regulating groundwater use: the challenges of policy implementation in Guanajuato, Central Mexico.** *Environ Sci Pol* 2017, **77**:107–113, <https://doi.org/10.1016/j.envsci.2017.08.002>.
29. Castellazzi P, Martel R, Rivera A, Huang J, Pavlic G, Calderhead AI, Chaussard E, Garfias J, Salas J: **Groundwater depletion in Central Mexico: use of GRACE and InSAR to support water resources management.** *Water Resour Res* 2016, **52**:5985–6003, <https://doi.org/10.1002/2015WR018211>.
30. *Plan Estatal de Desarrollo Guanajuato 2040. Guanajuato: gobierno del Estado de Guanajuato.* 2018. Available at, [https://seieg.iplaneg.net/seieg/doc/PED2040\\_sintesis\\_1525457699.pdf](https://seieg.iplaneg.net/seieg/doc/PED2040_sintesis_1525457699.pdf).
31. Schwarz J, Mathijs E: **Globalization and the sustainable exploitation of scarce groundwater in coastal Peru.** *J Clean Prod* 2017, **147**:231–241. Available at, <https://www.sciencedirect.com/science/article/pii/S0959652617300744?via%3Dihub>.
32. Samoral G, Carbó AV, Zegarra E, Knox JW: **Reconciling irrigation demands for agricultural expansion with environmental sustainability - a preliminary assessment for the Ica Valley.** *Peru. J Cleaner Prod* 2020, **276**:1–14, <https://doi.org/10.1016/j.jclepro.2020.123544>.
33. Damonte G, Domínguez I, Muñoz I: **Oré María Teresa: escasez de agua en la cuenca del río Ica y el Alto Pampas en Huancavelica. Un intento de mirada interdisciplinaria.** In *Aguas en disputa. Ica y Huancavelica, entre el entrapamiento y el diálogo*. Edited by Oré MT, Muñoz I; 2018. Available at, <https://www.untumbes.edu.pe/vcs/biblioteca/document/variostibros/0908.%20Aguas%20en%20disputa.%20Ica%20y%20Huancavelica.%20entre%20el%20entrapamiento%20y%20el%20di%20logo.pdf>; 2018.
34. *La superficie de riego eficiente en España se sitúa en 2.943.088 hectáreas, un 77 % de la superficie total de riego. Ministerio de Agricultura, Pesca y Alimentación.* 2021. Available at, [https://www.mapa.gob.es/es/prensa/210528superficieregadioenespana2020\\_tcm30-562657.pdf](https://www.mapa.gob.es/es/prensa/210528superficieregadioenespana2020_tcm30-562657.pdf).
35. *Encuesta sobre el uso del agua.* Instituto Nacional de Estadística; 2018. Available at, [https://www.ine.es/prensa/euasa\\_2018.pdf](https://www.ine.es/prensa/euasa_2018.pdf).
36. Paul C, Techen AK, Robinson JS, Helming K: **Rebound effects in agricultural land and soil management: review and analytical framework.** *J Clean Prod* 2019, **227**:1054–1067, <https://doi.org/10.1016/j.jclepro.2019.04.115>.
37. Song J, Guo Y, Wu P, Sun S: **The agricultural water rebound effect in China.** *Ecol Econ* 2018, **146**:497–506.
38. Fei R, Xie M, Wei X, Ma D: **Has the water rights system reform restrained the water rebound effect? Empirical analysis from China's agricultural sector.** *Agric Water Manag* 2021, **246**: 106690.
39. Berbel J, Gutiérrez-Martín C, Rodríguez-Díaz JA, Camacho E, Montesinos P: **Literature review on rebound effect of water saving measures and analysis of a Spanish case study.** *Water Resour Manag* 2015, **29**:663–678.
40. Wheeler SA, Carmody E, Grafton RQ, Kingsford RT, Zuo A: **The rebound effect on water extraction from subsidising irrigation infrastructure in Australia.** *Resour Conserv Recycl* 2020, **159**:104755.
41. Chindarkar N, Quentin G: **India's depleting groundwater: when science meets policy.** *Asia and the Pacific Policy Studies* 2018, **6**:108–124, <https://doi.org/10.1002/app5.269>.
- This is a well analysed paper on science and policy issues
42. *Acuíferos: gestión sostenible de las aguas subterráneas.* IUCN; 2020. Available at, <https://portals.iucn.org/library/node/49139>.
43. *Backgrounder: U.S. energy utilization for groundwater supply.* 2017. Available at, <https://www.ngwa.org/docs/default-source/>

- default-document-library/publications/information-briefs/energy-utilization-for-gw-background.pdf?sfvrsn=f5747a68\_2.
44. Cohen R, Wolff G, Nelson B: **Energy down the drain. The hidden costs of California's water supply** 2004. Available at, <https://www.nrdc.org/sites/default/files/edrain.pdf>; 2004.
  45. Hartung H, Pluschke L: *The benefits and risks of solar-powered irrigation - a global overview*. Rome: Food and Agriculture Organization; 2018. Available at, <https://www.fao.org/3/I9047EN/i9047en.pdf>.
  46. Kumar MD, Scott CA, Singh OP: **Can India raise agricultural productivity while reducing groundwater and energy use?** *Int J Water Resour Dev* 2013, **29**:557–573.
  47. Chen X, Thorp KR, Ouyang Z, Hou Y, Zhou B, Li Y: **Energy consumption due to groundwater pumping for irrigation in the North China Plain**. *Sci Total Environ* 2019, **669**:1033–1042. <https://www.sciencedirect.com/science/article/abs/pii/S0048969719311696>.
  48. Bennett B, Park L, Wilkinson R: *Embedded Energy in water studies: statewide and regional water--energy relationship*. California Public Utilities Commission; 2010. Available at, <http://www.waterenergyinnovations.com/publication/view/cpuc--embedded--energy--in--water--studies--1--statewide--and--regional--water--energy--relationship/>.  
A comprehensive study of embedded energy in water.
  49. Sayre SS, Taraz V: **Groundwater depletion in India: social losses from costly well deepening**. *J Environ Econ Manag* 2019, **93**:85–100. <https://doi.org/10.1016/j.jeem.2018.11.002>.
  50. Salinización de las aguas subterráneas en los acuíferos costeros mediterráneos e insulares españoles. Prepared by Custodio E., For UPC, SUEZ-Spain and cetaqua: 1–852. Technical University of Catalonia, Barcelona. Available at, <https://hdl-handle-net.ezproxy.lib.gla.ac.uk/2117/111515>.
  51. Bassi N: **Solarizing groundwater irrigation in India: a growing debate**. *Int J Water Resour Dev* 2018, **34**:132–145. <https://doi.org/10.1080/07900627.2017.1329137>.
  52. Oudra I, Talks P: *FAO/WB Cooperative Programme. Nationally determined contribution support on the groundwater, energy and food security nexus in Morocco*. 2019. Available at, <https://documents1.worldbank.org/curated/en/353851560191063136/pdf/FAO-WB-Cooperative-Programme-Nationally-Determined-Contribution-Support-on-the-Groundwater-Energy-and-Food-Security-Nexus-in-Morocco.pdf>.
  53. Refining estimates of water-related energy used in California. Navigant Consulting, Inc., for California Energy Commission. Available at, <https://calisphere.org/item/ark:/86086/n2hq3xr1/>.
  54. Hardy L, Garrido A, Juana L: **Evaluation of Spain's water-energy nexus**. *Int J Water Resour Dev* 2012, **28**:151–170. <https://www.tandfonline.com/doi/abs/10.1080/07900627.2012.642240>.
  55. Pienaar H, Xu Y, Braune E, Cao J, Dzikitii S, Jovanovic NZ: *Implementation of groundwater protection measures, particularly resource-directed measures in South Africa: a review paper*, vol. 23. Water Policy; 2021:819–837. <https://iwaponline.com/wp/article/23/4/819/82445/Implementation-of-groundwater-protection-measures>.
  56. Informe del Medio Ambiente. Gobierno de México, Secretaría de Medio Ambiente y Recursos Naturales. Available at, <https://apps1.semarnat.gob.mx:8443/dgeia/informe18/tema/cap6.html>.
  57. India's Atal Bhujal Yohana (Abhy) National Groundwater Management Improvement Programme. World Bank. Available at, <https://projects.worldbank.org/en/projects-operations/project-detail/P158119?lang=en>.
  58. China introduces first national plan on groundwater pollution control. Ministry of Water Resources, the People's Republic of China. Available at, [http://www.mwr.gov.cn/english/Medianews/201111/t20111108\\_309311.html](http://www.mwr.gov.cn/english/Medianews/201111/t20111108_309311.html).
  59. *Sustainable groundwater management Act (SGMA)*. California Department of Water Resources; 2021. Available at, <https://water.ca.gov/programs/groundwater-management/sgma-groundwater-management>.  
An excellent example of a long-term policy.
  60. Directive 2006/118/EC of the European Parliament and of the Council of 12 December 2006 on the protection of groundwater against pollution and deterioration. The European Parliament and the Council of the European Union. Available at: <https://eur-lex.europa.eu/legal-content/EN/ALL/?uri=CELEX:32006L0118>.
  61. Regulation (EU) 2020/741 of the European Parliament and of the Council of 25 May 2020 on minimum requirements for water reuse (Text with EEA relevance). The European Parliament and the Council of the European Union. Available at: <https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX%3A32020R0741>.
  62. *Policies to manage agricultural groundwater use: Australia*. Paris: Organisation for Economic Co-operation and Development; 2015. Available at, <https://www.oecd.org/australia/groundwater-country-note-AUS-2015%20final.pdf>.
  63. Horne J: **Australian water decision making: are politicians performing?** *Int J Water Resour Dev* 2020, **36**:462–483. <https://www.tandfonline.com/doi/abs/10.1080/07900627.2019.1685950?journalCode=cijw20>.
  64. *Agricultural policy monitoring and evaluation 2021*. Paris: Organisation for Economic Co-operation and Development; 2021. Available at, [https://www.oecd-ilibrary.org/agriculture-and-food/agricultural-policy-monitoring-and-evaluation\\_22217371](https://www.oecd-ilibrary.org/agriculture-and-food/agricultural-policy-monitoring-and-evaluation_22217371).
  65. Kehl J: **Moving beyond the mirage: water scarcity and agricultural use inefficiency in USA**. *Water* 2020, **12**:2290. <https://doi.org/10.3390/w12082290>.
  66. Turner SWD, Hejazi M, Calvin K, Kyle P, Kim S: **A pathway of global food supply adaptation in a world with increasingly constrained groundwater**. *Sci Total Environ* 2019, **673**:165–176. <https://doi.org/10.1016/j.scitotenv.2019.04.0700048-9697>.