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Food for contagion: Synthesis and future directions for studying host–parasite responses to resource shifts in anthropogenic environments

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Short title: Resource subsidy effects on disease

1 **Abstract.** Human-provided resource subsidies for wildlife are diverse, common, and have
2 profound consequences for wildlife–pathogen interactions, as demonstrated by papers in this
3 themed issue spanning empirical, theoretical, and management perspectives from a range of
4 study systems. Contributions cut across scales of organization, from the within-host dynamics of
5 immune function, to population-level impacts on parasite transmission, to landscape- and
6 regional-scale patterns of infection. In this concluding paper, we identify common threads and
7 key findings from author contributions, including the consequences of resource subsidies for (i)
8 host immunity; (ii) animal aggregation and contact rates; (iii) host movement and landscape-
9 level infection patterns; and (iv) inter-specific contacts and cross-species transmission. Exciting
10 avenues for future work include studies that integrate mechanistic modeling and empirical
11 approaches to better explore cross-scale processes, and experimental manipulations of food
12 resources to quantify host and pathogen responses. Work is also needed to examine evolutionary
13 responses to provisioning, and ask how diet-altered changes to the host microbiome influence
14 infection processes. Given the massive public health and conservation implications of
15 anthropogenic resource shifts, we end by underscoring the need for practical recommendations to
16 manage supplemental feeding practices, limit human–wildlife conflicts over shared food
17 resources, and reduce cross-species transmission risks, including to humans.

18

19

20 Key words: resource subsidy; anthropogenic change; human feeding of wildlife; pathogen
21 transmission; within-host dynamics; cross-species transmission

22 **Introduction**

23 Human feeding of wildlife is pervasive and can occur through both intentional (bird feeders,
24 tourist sites; [1,2]) and unintentional routes (landfills, agricultural crops; [3,4]). In response,
25 animal populations can shift movement behaviors or geographic ranges, experience higher
26 densities and contact rates, and show changes in demographic rates and interactions with other
27 species. The population- and community-ecological consequences of supplemental feeding in
28 wildlife have rarely been explored and could be far-reaching, particularly for infectious disease
29 dynamics [5–8].

30 Papers in this issue directly examine the interactions between anthropogenic resource
31 subsidy and infectious disease dynamics in wildlife using diverse approaches that include
32 mechanistic models, observational field studies and experiments, analysis of citizen science data,
33 and synthetic reviews. Empirical studies presented here examine diverse and engaging empirical
34 systems, ranging from birds at backyard feeders, to bats in urban and agricultural environments,
35 to elk in Yellowstone National Park (Figure 1). Despite differences in the biology of distinct
36 systems and environmental contexts, papers in this theme issue point to common questions,
37 patterns, and challenges for future work. Our goals in writing this synthesis are to identify these
38 common threads and outline several immediate priorities for future research on the links between
39 human resource subsidies and wildlife disease.

40 The taxonomic breadth of hosts and pathogens affected by resource provisioning, and the
41 range of food sources examined here, underscore how pervasive this phenomenon has become.
42 Given that responses of several pathogens studied here are accompanied by elevated risks of
43 cross-species transmission to humans, livestock, or vulnerable wildlife populations, studies that
44 provide a mechanistic understanding are sorely needed to predict future responses to feeding by

45 humans. The inevitability that human populations will continue to expand, alter habitats globally,
46 and encroach on wildlife, means that animal use of resources provided by humans will only
47 increase, lending a sense of urgency to understanding the impacts for wildlife, domestic animal,
48 and human health [9].

49

50 **Key findings and common threads across diverse systems and approaches**

51 *Host immunity shows complex responses to resource provisioning*

52 Because mounting and maintaining immune defenses require energy and nutrients [10,11],
53 access to anthropogenic food subsidies could increase the immune function of wildlife,
54 especially during times or in habitats where natural food sources are scarce or limited [12].

55 Under the common assumption that provisioning leads to better-defended hosts, pathogen
56 transmission should decrease owing to lower infection probability or faster recovery times
57 [13,14], but such effects might be offset by other processes like aggregation around food that
58 increase pathogen transmission [15]. Hite and Cressler (this issue) used a nested mechanistic
59 models to show that even if resources decrease host susceptibility to infection, an increase in host
60 densities in response to resource subsidies can override this effect and produce a higher total
61 transmission rate.

62 Empirical studies in this issue showed that the relationship between provisioning and
63 immunity can depend on the type of defense, quality of resources, and host and pathogen
64 taxonomy, leading to divergent outcomes among study systems (reviewed in Strandin et al., this
65 issue). This finding is consistent with past work on domesticated animals showing that different
66 components of host immunity respond differently to resource subsidies, in part because of the
67 variable costs of different immune process, and also because key macro- and micronutrients can

68 lead to immune system biases [16,17]. In natural systems, Becker et al. (this issue) found that
69 abundant livestock as food for vampire bats predicts stronger innate immunity relative to
70 adaptive immunity. Heightened innate immunity in the bats was further associated with a lower
71 probability of infection by *Bartonella* and hemoplasmas. Importantly, individual dietary history
72 itself did not strongly predict variation in bat immune profiles, suggesting that broader habitat-
73 level factors associated with livestock rearing could underlie parasite exposure and host
74 immunity. In other cases, such as elk supplemented at winter feedgrounds (Cotterill et al. (this
75 issue) and urban flying foxes [18], researchers hypothesized decreased immunocompetence with
76 food provisioning, owing to elevated stress hormones stemming from high host densities and due
77 to coinfections that impair immune response. Immune activity can also be compromised if
78 human-provided food is contaminated with toxins or drugs. As a case in point, Spanish imperial
79 eagles supplemented for conservation purposes with domestic rabbits (that had been treated with
80 antibiotics and antiparasitic drugs) showed decreased complement activity owing to the presence
81 of pharmaceuticals (especially fluoroquinolones) in their food [19]. Similarly, vampire bats that
82 fed more consistently on domestic animals in agricultural habitats had higher concentrations of
83 mercury that were associated with weaker bacterial killing ability of plasma [20].

84 It is important to note that evidence for nutritional condition altering wildlife immune
85 defenses is limited to a relatively small number of hosts, and studies of macro- and micro-
86 nutrient influences on immunity are needed to more critically evaluate this assumption. Genome-
87 wide RNA sequencing could help researchers focus on particular defense mechanisms by
88 quantifying immune gene expression between provisioned and unprovisioned groups, and those
89 with or without known infections [21,22]. In future work, phylogenetically informed meta-

90 analysis could help quantify the importance of food quantity, quality, and host and pathogen
91 traits [23] for immune defense and infection outcomes across wildlife systems.

92

93 *Behavioral changes in foraging and contact can alter local transmission processes*

94 Several studies in this theme issue demonstrate how resource provisioning can alter key
95 behaviors that underlie pathogen transmission, including foraging behavior, aggregation, and
96 contacts between species [15,24,25]. Crowding of individuals around supplemental resources can
97 lead to higher host densities and contact rates, and thus increase density-dependent transmission,
98 as illustrated previously through theoretical models [13]. Moyers et al. (this issue) designed an
99 experiment to test how feeder density influenced contact rates and exposure to the bacterium
100 *Mycoplasma gallisepticum* in captive house finches. Their work showed that higher bird feeder
101 density in enclosures caused the rapid spread of clinical infections, whereas lower feeder density
102 reduced pathogen spread, possibly due in part to the presence of sub-clinical and potentially
103 immunizing exposures. Importantly, further work is needed to examine how individual-level host
104 heterogeneity in the use of supplemental resources contributes to population-level infection
105 dynamics; for example, can subsets of hosts that aggregate around resources act as super-
106 spreaders, or might host heterogeneity limit the population-level spread of disease?

107 Cotterill et al. (this issue) reviewed the implications of intentional winter feeding of elk
108 (to limit encounters with cattle) in the western USA. Feed grounds have facilitated brucellosis
109 transmission among elk by elevating local density and contact rates [26] and, more speculatively,
110 by decreasing immune function. Feeding has now created a policy conundrum: high infection
111 prevalence in elk leads to greater motivation to separate elk and cattle, which leads to continued
112 winter feeding and further infection risk. While numerous papers in this theme issue advance a

113 mechanistic understanding of the links between disease and provisioning, disentangling the roles
114 of aggregation and subsequent contact rates, versus changes in immune functions, for driving
115 pathogen transmission will require further work.

116 Resource provisioning often causes changes in diet and foraging behaviors, especially
117 among urbanized wildlife populations that subsist on supplemental food. Murray et al. (this
118 issue) showed that white ibises shifting from natural wetlands to urban parks in Florida, where
119 they commonly forage on provisioned food, have lower ectoparasite burdens. To explain this
120 pattern, the authors hypothesize that easier food access might allow birds to spend less time
121 foraging and more time preening to remove parasites. In urban and coastal Queensland, the
122 Australian white ibis experienced explosive population growth in the 1990s due to provisioning
123 from open landfills [27,28]. The abundance of anthropogenic food waste as well as deliberate
124 feeding in urban parks led to a shift from coastal nesting and foraging to suburban and urban
125 foraging, bring ibis into greater contact with each other, as well as with chickens on poultry
126 farms and people in recreational areas [27]. Increased population density and interaction among
127 ibis and with domesticated animals and people could also increase risk of intra- and interspecies
128 pathogen transmission. Understanding the mechanistic links between shifts in behavior and
129 disease risk could be strengthened by future studies that simultaneously measure specific
130 behaviors (at the individual level) and changes in infection (at individual and population levels).
131 For some food-provisioned populations, efforts to limit contact rates during high-risk intervals
132 (e.g., by ending feed dates earlier in the season for elk, or spacing out bird feeders at lower
133 density) or preserve particular behaviors (e.g., such as preening or other anti-parasite behaviors)
134 could prove important for managing infection risk in wildlife.

135

136 *Behavioral changes in host movement can influence landscape-level disease processes*
137 Provisioning can cause changes to host movements and infection patterns at large spatial scales.
138 As reviewed by Satterfield et al. (this issue), anthropogenic food subsidies can decrease
139 migratory movements and concentrate hosts into resource-subsidized regions, where greater host
140 aggregation, year-round parasite accumulation, and longer residency times could increase
141 exposure to pathogens [6,23]. The authors note that shifts towards more sedentary behavior in
142 response to resource provisioning has occurred for multiple migratory and nomadic species, in
143 some cases associated with resulting increases in infection risk [29–32]. For example, satellite
144 telemetry studies of *Pteropus medius*, the reservoir for Nipah virus in Bangladesh, suggest that
145 this species is much more sedentary than its relative, *P. vampyrus* in Malaysia, which could be
146 due, in part, to anthropogenic food resources (Epstein et al., *unpublished*) [33]. Date palm sap,
147 harvested by humans in Bangladesh, is exploited by frugivorous bats throughout winter months
148 and is the primary route of Nipah virus spillover from bats to people [34,35]. Alternatively,
149 animals that stop migrating might be exposed to a lower diversity of parasites across their
150 migratory range, and more limited host movements could reduce the spatial spread of pathogens
151 [36,37]. A theoretical model (Brown and Hall, this issue) explored these questions for a partially
152 migratory host affected by a vector-borne pathogen. The model showed that when provisioning
153 increased the survival of resident hosts during the non-breeding season, both infection
154 prevalence and the fraction of the population that is non-migratory increased. Because greater
155 proportions of residents permit the sustained transmission of pathogens, this behavioral shift
156 could be especially costly to remaining migrants that travel through areas with infected residents;
157 resource provisioning could therefore threaten the persistence of migratory behavior.

158 For some highly mobile hosts, resource provisioning will alter daily foraging movements
159 and habitat use. In Australia, naturally nomadic fruit bats have shifted into urban areas where
160 they feed on native and exotic flowering and fruiting trees planted by humans [30,38]. Paez et al.
161 (this issue) applied optimal foraging theory to explore how urban bat colonies alter their
162 foraging strategies in response to decreasing native habitat and seasonal food availability. Their
163 work predicts that residency in urban patches will increase as native foraging habitats become
164 more isolated, and during periods of overall food scarcity. Longer residency in urban centers
165 could set the stage for less frequent but larger viral outbreaks in bats, resulting in higher
166 exposure to humans and domesticated animals [38,39].

167

168 *Changes to interspecific interactions can cause cross-species transmission and pathogen*
169 *emergence*

170 Cross-species pathogen transmission requires several ecological, epidemiological, and behavioral
171 factors to align [40]. Importantly, anthropogenic provisioning can influence multiple components
172 of this alignment by (i) changing host community composition, (ii) altering infection dynamics
173 within populations of reservoir hosts, and (iii) affecting contact rates between host species. First,
174 because the responses of host species to novel resources in human-altered landscapes can range
175 from disappearance to explosive population growth, provisioning can dramatically alter host
176 community composition and patterns of pathogen transmission [41–43]. As an example of these
177 changes, large-scale monocultures in Brazil and Panama altered rodent communities and
178 increased human exposures to rodent species infected with hantavirus [44,45]. At the largest
179 spatial scales, provisioning could expand host geographic ranges, creating novel opportunities
180 for cross-species transmission where hosts previously did not co-occur [46]. Second, changes to

181 infection dynamics within primary host species (see above sections) can have knock-on effects
182 that amplify or dampen the probability of transmission given inter-specific contacts [15]. Third,
183 even if host community composition and disease dynamics in reservoir species remain
184 unchanged, provisioning can facilitate cross-species transmission by altering the frequency and
185 nature of inter-species contacts. For example, bats foraging on mango trees planted near pig
186 farms, or bats drinking palm sap as it runs down tree trunks into collecting vessels, created new
187 routes of Nipah virus transmission from bats to pigs and humans, respectively [35,47]. The
188 common practice of allowing domestic animals to feed on dropped or bitten fruit, that may have
189 been contaminated by bats, also increases the risk of pathogen transmission [48,49]. In
190 Bangladesh, 26 common fruits grown and eaten by people are known to be eaten by frugivorous
191 bats, and eating dropped fruit with animal bite marks regularly occurs (Epstein et al.,
192 unpublished.) Similar processes could influence pathogen transmission among wildlife when
193 resources promote multi-species aggregations of previously ecologically isolated species [50,51].
194 Importantly, these mechanisms of resource-driven changes in cross-species transmission might
195 act synergistically. As discussed by Becker et al. (this issue), livestock both stimulates vampire
196 bat population growth and, by its own presence, expands opportunities for cross-species
197 transmission of rabies virus and potentially other pathogens.

198 Altered dynamics of cross-species transmission are among the most visible and alarming
199 responses to resource provisioning because they can directly impact human health, agriculture, or
200 the conservation of vulnerable wildlife populations. For example, livestock-driven increases in
201 vampire bat rabies have made this disease one of the three most important zoonoses in Latin
202 America and a significant barrier to the advancement of agrarian communities [52,53]. Similarly,
203 the resource-driven rise of Hendra virus cases in humans and horses in Australia created

204 economic and social challenges, ranging from the rising need for veterinary vaccines to protect
205 horses, to conservation challenges as bat persecution is promoted for disease control [38,54]. In
206 Asia, the transmission of zoonoses from provisioned non-human primates to people impacts
207 tourism [55]. Importantly, provisioned landscapes can provide opportunities for spillover
208 infections from humans (or livestock) to wildlife, and potential spillback into humans. For
209 example, in parts of Africa, baboons commonly frequent human settlements and obtain food
210 from houses or waste sites. Parasitological surveys showed baboons near these settlements can
211 harbor parasitic worms and protozoa that commonly infect humans, although further diagnostic
212 work is needed to determine whether the primate isolates match parasite genotypes recovered
213 from nearby humans [56,57]. Better quantifying the contexts under which provisioning mediates
214 cross-species transmission could provide an epidemiological lever to promote more responsible
215 management of anthropogenic food subsidies for wildlife.

216

217 **Critical priorities for future work**

218 *Taxonomic biases in studies of provisioning and infection*

219 Work included in this Theme Issue reflects the taxonomic breadth of hosts and parasites studied
220 in the context of resource provisioning, and also highlights taxonomic gaps to be addressed in
221 future work. The empirical studies presented here focus primarily on mammals (e.g., bats,
222 ungulates) and birds (e.g., passerines, wading birds), with less representation from invertebrates
223 (e.g., monarch butterflies, *Daphnia*). Studies here also focused heavily on microparasites,
224 particularly bacteria and viruses, transmitted through direct and non-close contact (e.g., fecal-
225 oral routes), although ectoparasites are also represented. More generally, throughout the
226 literature, studies of provisioning and host-parasite interactions are biased towards these taxa

227 (reviewed in [15, 23]). For example, a recent meta-analysis of over 300 host-parasite interactions
228 was dominated by studies of microparasites transmitted by close and non-close contact, and of
229 helminths transmitted through non-close contact and intermediate hosts [23]. Vector-borne
230 diseases, and protozoan and fungal parasites are generally poorly represented, highlighting a
231 priority for future studies, particularly in light of expanding vector distributions under climate
232 change and the role of fungal parasites in wildlife population declines [58–61]. Past studies of
233 food provisioning and wildlife disease also heavily biased towards mammals and birds, with
234 much less work on invertebrates and other ectotherms. Civitello et al. (this issue) highlight how
235 nutrient inputs into aquatic ecosystems (as a form of anthropogenic subsidy) can have similar
236 effects on host-parasite interactions as food subsidy to wildlife (by increasing host density and
237 altering parasite production within hosts). This observation stresses the need for greater inclusion
238 of amphibians, reptiles, fish, and invertebrates in studies of provisioning and disease.

239

240 *Modeling studies to link effects of provisioning across biological scales*

241 Resources can affect within-host processes relevant to pathogen colonization, between-host
242 transmission at the population level, and landscape-level processes such as host dispersal.
243 Mathematical models provide powerful tools for linking infection dynamics across scales of
244 organization and for informing the conditions under which provisioning can increase or decrease
245 infection. For example, theory to date has shown that when resources strongly enhance host
246 defenses, this can limit pathogen transmission that otherwise would increase from resource-
247 induced increases in host density [13,62]. If host defenses are unchanged or weakened by
248 human-provided resources, increased exposure to pathogens resulting from elevated host
249 densities and behavioral changes are likely to increase pathogen invasion and prevalence [15]. A

250 separate body of theory used metapopulation models to examine how the distribution of
251 resource-rich habitats, and their impact on colonization and extinction, affects host-pathogen
252 dynamics. This work shows that increasing the frequency of provisioning across the landscape
253 increases pathogen establishment and spread; yet nonlinear relationships between infection
254 prevalence and the relative abundance of provisioned habitats can emerge if provisioning and
255 infection influence host movement decisions and dispersal success [63,64]. Despite these recent
256 advances, a need remains for mathematical models that more explicitly link processes across
257 individual, population, and landscape scales.

258 In this issue, Hite and Cressler contribute a cross-scale approach by developing a
259 mechanistic framework coupling within-host processes (through improved immune defense and
260 increased pathogen replication in response to resources) and between-host processes (through
261 transmission and resource-mediated population growth rates). Their model explores the
262 consequences of resource acquisition for parasite virulence evolution and its potential to stabilize
263 resource-driven cycles in host population dynamics. The authors demonstrate that linking within-
264 host and population-level processes can produce cyclic host population dynamics and associated
265 within-host cycles of high and low parasite replication, an emergent phenomenon that does not
266 occur when within-host processes are ignored. In other work, Civitello et al. (this issue)
267 demonstrate that incorporating trophic complexity (by considering predators and competitors of
268 provisioned hosts) can reverse predictions about resource-mediated increases in pathogen
269 prevalence. Resource subsidies increase pathogen prevalence when only hosts are present, but
270 competitors and predators can lower infection prevalence (in some cases causing pathogen
271 extinction) when resources are abundant. These studies highlight the importance of considering
272 processes at scales above and below the population level in predicting resource subsidy effects

273 on pathogen transmission dynamics. An additional key insight from theoretical work is that
274 empirical studies must be long enough relative to the duration of infection to capture stable or
275 cyclic responses of population and infection dynamics under provisioning. Promising future
276 avenues include investigating how resources affect coinfection (e.g., in shaping immune-
277 mediated competitive interactions between micro- and macroparasites); the responses of
278 parasites with complex transmission modes (e.g., vector-borne and trophically transmitted
279 parasites); and relationships for multi-host pathogens where host species that differ in
280 competence might respond differently to provisioned resources (e.g., in population density or
281 susceptibility to infection) [9].

282 Future theoretical models that are paired closely with detailed empirical work could be
283 especially fruitful in understanding the dynamical outcomes of provisioning. Such work could
284 couple local and landscape-level effects of resources on well studied host-pathogen interactions.
285 Given that theory to date on provisioning and infection has focused separately on population and
286 metapopulation scales, one area that is crucially needed involves models that explicitly link local
287 dynamics (e.g., resource effects on individual hosts or contact rates) to regional movements of
288 the host and pathogen that also depend on resource distributions (Figure 2). From an applied
289 perspective, such models could also allow researchers to predict the outcomes of different habitat
290 management scenarios that might alter resources in ways that lower infection risks [13,63,65].

291

292 *Experimental manipulations of food resources to quantify responses of hosts and pathogens*

293 Research manipulating food resources is noticeably rare among the growing body of literature
294 developing around the effects of anthropogenic food subsidies on host-parasite dynamics.
295 Indeed, this theme issue reflects this disparity between observational and experimental

296 approaches, with only a single study (Moyers et al., this issue) among the latter. A handful of
297 studies published elsewhere have experimentally manipulated food; for example, work by
298 Wright and Gompper [66] showed that clumped food resources increased the transmission of
299 endoparasites in raccoons, suggesting a possible behavioral mechanism for changes in
300 prevalence. Wilcoxon et al. [67] and Galbraith et al. [68] both manipulated the presence or
301 absence of bird feeders and found effects of feeder presence on health-associated traits such as
302 body condition, as well as effects on the prevalence of diverse parasites and pathogens.
303 Responses to feeder presence in Galbraith et al. [68] were parasite- and host-specific. Although
304 experimental in nature, field studies such as these still have difficulty establishing definite
305 causation (e.g., in contrast, see [69]). For example, in some systems, diseased animals could be
306 more strongly motivated to seek out supplemental food resources, leading to patterns of higher
307 infection prevalence at supplemented sites that could also be interpreted as a positive effect of
308 resources on pathogen transmission [70].

309 Most experiments to date manipulate food through experimental supplementation, but
310 future work could reduce access to anthropogenic foods, especially for species for which finding
311 or monitoring unprovisioned populations is difficult. For example, vampire bats in Latin
312 America are most abundant and thus readily sampled near livestock-rich areas [71], and locating
313 unprovisioned rainforest populations is difficult [72]. Moreover, multiple confounding factors,
314 including habitat characteristics and host density, differ between provisioned and unprovisioned
315 groups (Becker et al., this issue). In this case, restricting access to livestock, such as through
316 artificial lighting to deter bat feeding [73], might be one way to monitor host and pathogen
317 responses to reduced access to anthropogenic food. For other hosts, limiting access to human

318 foods through fencing, or through campaigns to restrict tourist feeding of wildlife, could generate
319 heterogeneity in resources.

320 Manipulating food quantity and quality is needed to explore the effects of food nutritional
321 value on multiple measures of host immune defense, the host microbiome (discussed below), and
322 susceptibility to target pathogens. Some experimental provisioning work has examined
323 individual and population-level outcomes in birds and rodents [74–77]. Many of these
324 experiments have been conducted in semi-controlled settings, such as aviaries and field
325 enclosures, reflecting challenges associated with regulating food and disease exposure in free-
326 ranging wildlife, which can disperse over large areas. However, confinement might also impact
327 disease outcomes in unnatural ways, such as by increasing the frequency and intensity of
328 intraspecific transmission opportunities, and inducing stress that often impairs host immunity
329 (Strandin et al., this issue).

330 Future field experiments might simultaneously control multiple components of
331 provisioning, especially if anthropogenic foods dampen the seasonality or pulsed timing of
332 natural resources, and at the same time make food more spatially aggregated, or change resource
333 quality. These same studies could experimentally reduce infections in some hosts, to separate
334 responses of host behavior, physiology and fitness from parasite infection itself. Given the
335 pervasiveness of provisioning, many opportunities exist to integrate experiments within current
336 feeding activities, particularly within wildlife management and conservation efforts (e.g.,
337 Cotterill et al., this issue). Moreover, the strong causal inference provided by well-planned and
338 executed experiments (e.g., by manipulating both infection and resources in free-ranging wildlife
339 [74]) necessitates greater emphasis on these approaches to better understand how anthropogenic
340 resources affect host–parasite dynamics.

341

342 *Understanding consequences of resource subsidies for the evolution of pathogen virulence*

343 By affecting pathogen transmission and within-host processes, resource provisioning could
344 ultimately affect host and pathogen evolution, an idea explored in depth by Hite and Cressler
345 (this issue). General theory on virulence evolution predicts that greater opportunities for
346 horizontal pathogen transmission, such as might be created by aggregation around provisioned
347 resources, could favor the evolution of more virulent pathogen strains [78]. As described earlier,
348 Hite and Cressler's paper used a multi-scale model to show that such a result can arise even
349 when provisioning increases host immunity. Empirical work is crucially needed from naturally-
350 occurring host-pathogen systems to test the virulence of pathogen strains from provisioned and
351 unprovisioned host populations (e.g., [29]).

352 Although not examined by papers in this issue, provisioning can, in some cases, allow
353 wildlife to better tolerate infection [15], an idea supported by laboratory studies demonstrating
354 that improved nutrition can prolong the survival of infected animals and increase the duration of
355 pathogen shedding [79,80]. Because host mortality cuts short the infectious period for many
356 pathogens, this can constrain greater within-host replication by pathogens, and hence limit
357 virulence evolution. In contrast, more tolerant hosts could select for more virulent pathogen
358 strains by releasing pathogens from some of the costs of virulence [79]. Thus, although improved
359 condition could reduce disease-induced mortality of provisioned hosts in the short-term,
360 provisioning could favor the evolution of higher virulence in the longer term [81]. Evolutionary
361 models and empirical studies that explore the impact of resource subsidies on host tolerance to
362 infection, within the context of other processes, are needed to identify the conditions under
363 which provisioned populations support pathogen strains of higher virulence.

364

365 *Seeking how changes to the host microbiome affect larger-scale infection processes*

366 Another important area for future work is understanding how dietary changes associated with
367 provisioning could impact the host microbiome and within-host dynamics [15]. The composition
368 of gut microbial communities can influence the immune system, thereby affecting host
369 susceptibility and pathogen colonization [82]. For example, experimental simplification of
370 microbiota from Cuban tree frog tadpoles increased their susceptibility to invasion by gut
371 helminths as adults [83]. The composition and diversity of the gut microbiome is itself strongly
372 shaped by individual diet [84,85], and thus provisioned wildlife would be expected to differ in
373 both their microbiota and their susceptibility to enteric pathogens. Yet field studies of
374 microbiomes in provisioned hosts are rare; in one example, the gut microbiota of baboons
375 foraging on leftover food in Bedouin settlements mirrored the gut microbiota of people living in
376 the Bedouin communities [86].

377 Comparative work on the microbiome between provisioned and wild populations is
378 necessary to establish how specific dietary differences influence gut microbial composition and
379 diversity. For example, shifts from protein- to carbohydrate-rich diets in urban-foraging wildlife
380 such as white ibis (Murray et al., this issue) could have especially pronounced effects on
381 microbiomes, and, in turn, pathogen invasion. In one rare case study, shifts toward grain-based
382 diets may have disrupted the microbiota of Canada geese and facilitated *Clostridium perfringens*
383 colonization [87]. From another perspective, foraging on anthropogenic resources in urban and
384 agricultural environments could also expose species such as vampire bats and flying foxes (Paez
385 et al., this issue; Becker et al., this issue) to contaminants (e.g., pesticides and antibiotics) that
386 alter microbial community composition [88]. When possible, manipulative experiments are

387 needed to examine causal relationships between different components of provisioned diets and
388 the microbiome. Moreover, relationships among microbiome diversity, microbiome composition,
389 and susceptibility to pathogen challenge in the context of provisioning must be elucidated to
390 understand how changing microbiota influences host susceptibility to infection. Finally, data
391 linking diet, microbial diversity, and immunity could be used to parameterize mathematical
392 models to holistically explore how provisioning influences infection dynamics.

393

394 **Implications of provisioning for conservation and human health**

395 *The importance of understanding human motivations for feeding wildlife*

396 The pervasiveness and popularity of intentional wildlife provisioning (e.g., Cox and Gaston, this
397 issue) suggests that humans have strong underlying motivations for this activity, particularly in
398 the case of backyard bird feeding, on which people spend \$4.5 billion annually in the U.S. alone
399 [89]. Although bird feeding is the most prevalent form of intentional provisioning, a clear picture
400 of the disease risks this activity imposes on wildlife and humans remains elusive [70]. The
401 intentional feeding of charismatic mammals is common and probably alters disease risk as well.
402 For example, provisioning of wild primates is prevalent within the context of Hindu and
403 Buddhist culture, and has been enhanced with increasing tourism [90]. Motivations for feeding
404 wildlife are complex and may vary regionally [91,92], but numerous studies have shown a key
405 impetus of the psychological benefits of direct human-wildlife interaction [93], including a sense
406 of pleasure or relaxation, feelings of usefulness, and an increased connection to nature [94–97].
407 In fact, the vast majority of people surveyed about their willingness to interact with wild
408 primates were aware of the potential disease risks associated with this interaction, and yet more
409 than half still responded that they would touch wild primates if given the opportunity [98].

410 Welfare motivations are also commonly cited by those who provision wildlife [96], including a
411 desire to help wildlife or “assist them through hard times” [95,97]. Indeed, provisioning tends to
412 be strongest in seasons when natural food is perceived to be limited [95], suggesting a strong role
413 of welfare motivations.

414 Cox and Gaston (this issue) suggest that positive reactions from wildlife, as well as
415 psychological benefits to humans, strongly motivate people to offer supplemental foods,
416 although more empirical evidence is needed. For example, humans that receive significant
417 positive benefits from feeding (increased well-being or reduced stress) are probably more likely
418 to continue provisioning. On the other hand, Cox and Gaston (this issue) also propose that the
419 negative consequences of supplemental feeding, such as disease transmission among wildlife
420 (Lawson et al., this issue), or human health risks, often do not feed back to dampen provisioning
421 behavior because these effects are rarely apparent to the public [95]. The recent trend toward
422 reduced feeding of (non-bird) wildlife in the United States (Cox and Gaston, this issue) suggests
423 that active campaigns against feeding of mammals are beginning to influence human behavior.
424 Thus, by tapping into the welfare motivations for feeding wildlife, changes in human behavior
425 are possible. Success in changing behavior might be more even more likely when campaigns
426 directly target the negative effects on humans, such as in cases of human-wildlife conflict and
427 pathogen spillover.

428 To the extent possible, intentional supplemental feeding should be managed to maximize
429 benefits to both humans and wildlife. For example, the recently documented association between
430 higher levels of afternoon bird abundance and reductions in the severity of depression, anxiety,
431 and stress in humans led the authors to propose the active use of supplemental feeding to create
432 “optimal” bird abundance levels for human health [99]. For many bird species, supplemental

433 feeding decreases starvation risk [100] and can improve breeding success [101]. Yet, feeding has
434 also been associated with changes in community structure [102], range expansion [103], and, as
435 this issue illustrates, pathogen transmission. Unfortunately, it seems unlikely that optimal levels
436 of feeding for humans and wildlife will coincide. Thus, given the species- and habitat-specific
437 effects of supplemental feeding [102,104], determining the ideal levels of provisioning for most
438 wildlife will be challenging. In cases where clear negative effects of resource provisioning on
439 wildlife are documented, educational campaigns would ideally leverage welfare-driven
440 motivations for feeding by creating negative feedback loops on human behavior (Cox and
441 Gaston, this issue). Overall, effective management of intentional provisioning will require
442 significantly more data than are currently available on both human motivations for feeding,
443 effects of feeding on wildlife, and potential feedback loops between wildlife effects and human
444 behavior. Given the enormous and potentially growing scale of human supplementation of
445 wildlife [89], developing effective management tools is both timely and critical.

446

447 *Recommendations for limiting disease risks associated with human–wildlife contacts*

448 The proximity with wildlife afforded by resource subsidies in urban and agricultural landscapes
449 brings humans and domestic animals into contact with wildlife pathogens, and wildlife into
450 contact with human pathogens (Figure 3). Some of the most readily observed examples include
451 growing populations of urban mesocarnivores (e.g., foxes, raccoons and skunks) that can attack
452 humans and domestic animals when infected with rabies [105]. Non-human primates can also
453 become aggressive following habituation to human-provided food, leading to the transmission of
454 zoonotic viruses in some cases [106], and exposing primates to respiratory infections from
455 human researchers and tourists in other scenarios [107]. Wildlife professionals might be exposed

456 to zoonotic pathogens when translocating nonhuman primates in response to human-wildlife
457 conflict [108]. Even when interspecific contacts between wildlife and humans are rare,
458 pathogens can transfer between humans and wildlife by environmental routes or through
459 arthropod vectors. Examples include a rise in human infections with the soil-borne tapeworm
460 *Echinococcus multilocularis*, attributed to provisioned urban red foxes in Europe [109]. Human
461 and animal Nipah virus infections have occurred through the consumption of food contaminated
462 by bat excreta [110,111], and greater human exposures to hantavirus through environmental
463 infectious stages followed the growth of rodent populations that exploit agricultural crops
464 [44,45]. Zooanthroponoses (pathogens transmitted from humans to other animals) are less
465 appreciated, but affect wildlife globally [112,113]. The preponderance of environmentally and
466 vector-transmitted pathogens at the human-wildlife interface raises important challenges to
467 recognize links with resource provisioning. Epidemiological investigations that identify agents of
468 disease must be followed with ecological studies to identify natural hosts and the ecological
469 context that enables cross-species transmission [114]. Fortunately, rapid and powerful
470 DNA/RNA sequencing technologies [115], together with increasingly sophisticated tools for
471 inferring pathogen transmission between species [116] offer currently under-utilized
472 opportunities to improve scientific understanding of the changing patterns of pathogen
473 transmission in provisioned environments.

474 Under some circumstances, ecological interventions that build on a mechanistic
475 understanding of host and pathogen biology can prevent cross-species transmission. Most
476 notably, preventing wildlife access to unintentionally provisioned resources, or creating a barrier
477 between provisioned resources and domesticated animals (e.g., planting orchards away from
478 livestock enclosures to reduce the risk of Nipah spillover on farms in Malaysia), can restrict

479 opportunities for overlap between host species and function as a barrier to pathogen spillover
480 [117]. As one key example, blocking the foodborne transmission of Nipah virus from pteropid
481 fruit bats to humans using a bamboo skirt placed at the top of date palm sap collection pots
482 restricts bat access to this shared food resource, and could reduce the risk of Nipah virus
483 exposure in humans [118,119]. This case study highlights not only how basic ecological data on
484 the foraging behavior of reservoir hosts can aid in the design of interventions, but also how
485 insights from social science and the application of locally available practices can produce
486 economically affordable management tools [120]. Such “ecological interventions” may also be
487 cheaper and more effective than antibiotics or vaccines that are mobilized after cross-species
488 exposures occur. Other intervention strategies can promote sanitary best practices to prevent the
489 build-up on infectious stages on feeders (e.g., washing backyard bird feeders), encouraging the
490 dispersal of feed in smaller units over larger areas to reduce aggregation and lower contact rates
491 (e.g., with management-based feeding [121]), and educating the public about disease risks posed
492 by well-intentioned but harmful feeding activities [122,123]. Given that resource provisioning is
493 ultimately derived from human actions, perceptions, and policies, the integration of ecological,
494 sociological, and management perspectives will be a key lever by which infectious disease risks
495 can be minimized for the well-being of humans, domesticated animals, and wildlife.

496

497 **Competing interests**

498 The authors have no competing interests associated with this manuscript.

499 **Authors' contributions**

500 SA and DJB developed and organized the paper, drafted sections of the paper, integrated author
501 contributions, and created associated figures. All other authors participated in the design of the
502 paper, drafted sections of the manuscript, and helped revise the paper. All authors gave final
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504

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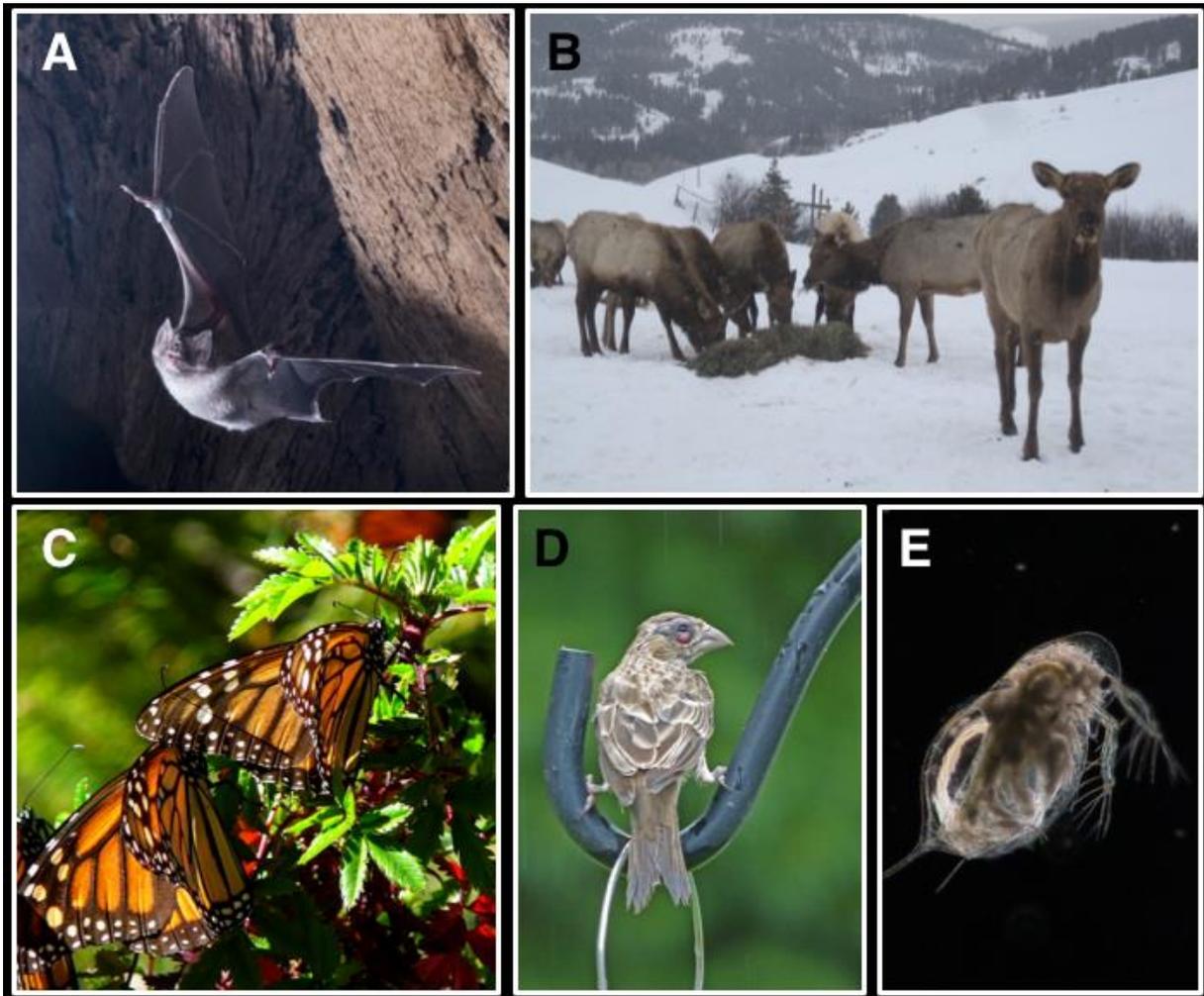
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836 **Figures**

837 Figure 1. Taxonomic breadth of hosts provisioned by humans covered by studies in this Theme
838 Issue: A. Common vampire bat (*Desmodus rotundus*) in Belize (Brock Fenton), B. Elk (*Cervus*
839 *elaphus*) in the Greater Yellowstone Ecosystem (Paul Cross), C. Monarch butterflies (*Danaus*
840 *plexippus*) in Mexico (Natalie Tarpein), D. House finch (*Haemorhous mexicanus*) infected with
841 *Mycoplasma gallisepticum* in North America (Bob Vuxinic), and E. *Daphnia dentifera* infected
842 with a fungal pathogen (*Metschnikowia bicuspidata*) (Tad Dallas [124]).

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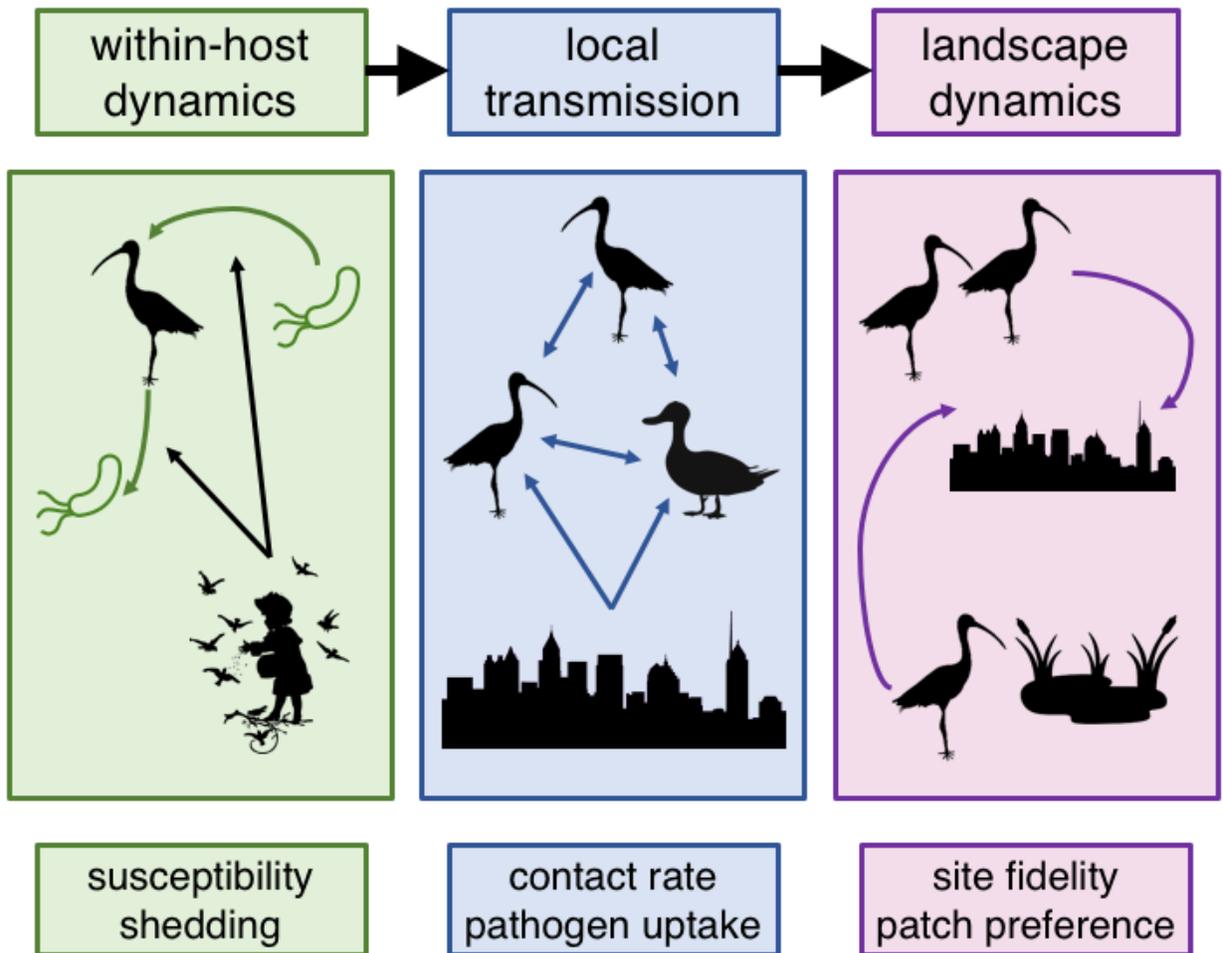


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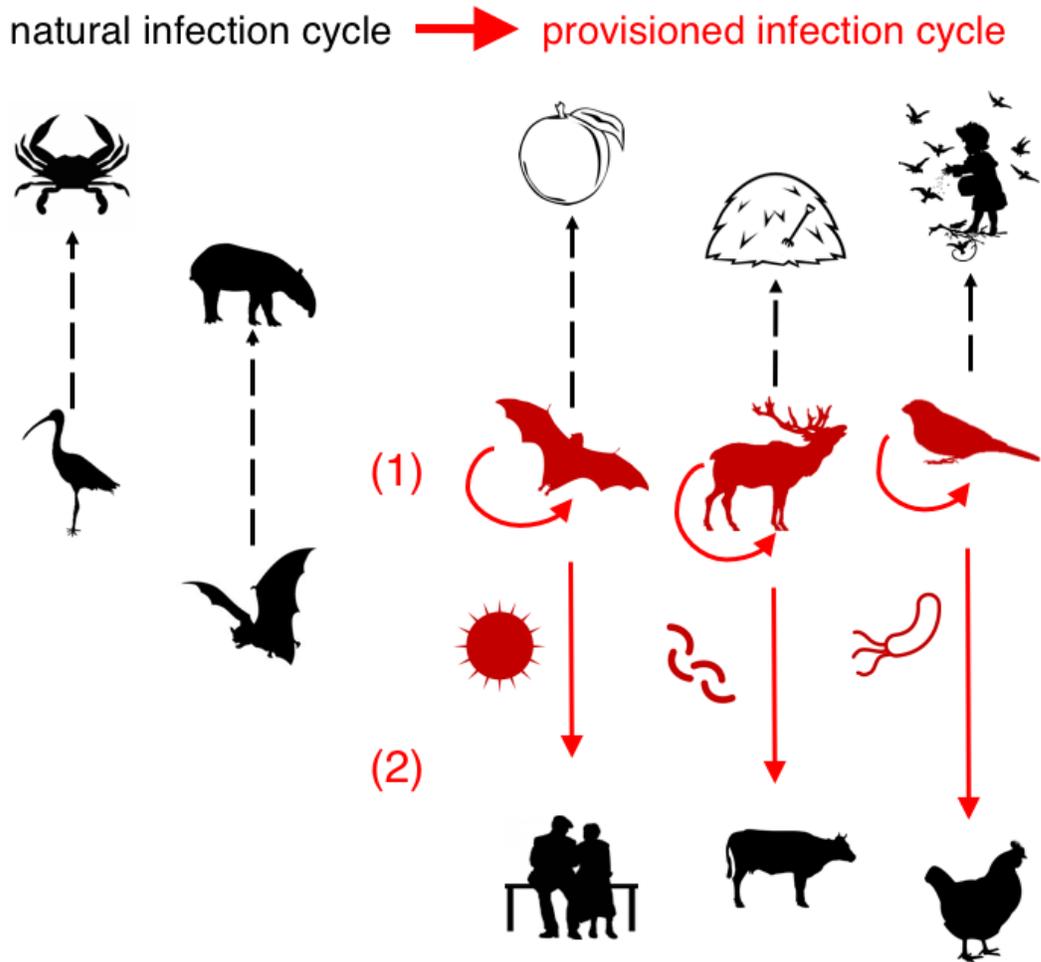
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847 Figure 2. Interactions between human-provided food and pathogen dynamics can occur at
 848 multiple scales of organization, as illustrated by American white ibis (*Eudocimus albus*) and
 849 environmentally-transmitted enteric pathogens. Anthropogenic food subsidies in urban habitats
 850 could influence within-host dynamics (e.g., individual susceptibility and intensity of pathogen
 851 shedding, in green), local transmission processes (e.g., intra- and inter-specific contact rates,
 852 uptake of pathogen from the environment, in blue), and landscape dynamics (e.g., host
 853 movement between natural and provisioned habitats, site fidelity, in pink). Combined modeling
 854 and empirical work is needed to quantify the importance of processes operating within scales,
 855 and to predict how processes at one scale affect dynamics at larger scales of organization.



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858 Figure 3. Possible effects of provisioning on amplifying pathogen spillover risks by 1) increasing
 859 pathogen transmission and shedding from reservoir hosts (e.g., through increased aggregation,
 860 susceptibility, and shedding intensity) and 2) increasing opportunities for contact between
 861 humans and domestic animals and either reservoir hosts or pathogen in the environment.
 862 Silhouettes and arrows display case studies from this theme issue where provisioning had little
 863 effect or decreased infection relative to more natural environments (black; white ibis, vampire
 864 bats) and where provisioning amplified infection cycles (red; flying foxes, elk, house finches)
 865 and could potentially increase the risks of cross-species transmission.
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