

Clements, E.A., Thomas, R. and Adams, C.E. (2018) An investigation of salmonid host utilisation by the endangered freshwater pearl mussel (*Margaritifera margaritifera*) in north-west Scotland. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 28(3), pp. 764-768.

There may be differences between this version and the published version. You are advised to consult the publisher's version if you wish to cite from it.

Clements, E.A., Thomas, R. and Adams, C.E. (2018) An investigation of salmonid host utilisation by the endangered freshwater pearl mussel (*Margaritifera margaritifera*) in north-west Scotland. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 28(3), pp. 764-768.
(doi:[10.1002/aqc.2900](https://doi.org/10.1002/aqc.2900))

This article may be used for non-commercial purposes in accordance with [Wiley Terms and Conditions for Self-Archiving](#).

<http://eprints.gla.ac.uk/157022/>

Deposited on: 09 February 2018



An Investigation of Salmonid Host Utilisation by the Endangered Freshwater Pearl Mussel (*Margaritifera margaritifera*) in north-west Scotland

Journal:	<i>Aquatic Conservation: Marine and Freshwater Ecosystems</i>
Manuscript ID	AQC-17-0165.R1
Wiley - Manuscript type:	Short Communication
Date Submitted by the Author:	n/a
Complete List of Authors:	Clements, Elizabeth; Envirocentre Thomas, Rhian; University of Glasgow, School of Geographical and Earth Sciences Adams, Colin; University of Glasgow, Scottish Centre for Ecology and the Natural Environment
Broad habitat type (mandatory) select 1-2:	river < Broad habitat type, stream < Broad habitat type
General theme or application (mandatory) select 1-2:	behaviour < General theme or application, endangered species < General theme or application
Broad taxonomic group or category (mandatory, if relevant to paper) select 1-2:	invertebrates < Broad taxonomic group or category
Impact category (mandatory, if relevant to paper) select 1-2:	

SCHOLARONE™
Manuscripts

1
2
3 **1 An Investigation of Salmonid Host Utilisation by the Endangered**
4 **2 Freshwater Pearl Mussel (*Margaritifera margaritifera*) in north-west**
5 **3 Scotland**
6
7
8

9
10 4 E.A. CLEMENTS^a, R. THOMAS^{b*} and C.E. ADAMS^a

11
12 ^aScottish Centre for Ecology and the Natural Environment, University of Glasgow,
13 Glasgow, U.K.
14

15
16 ^bSchool of Geographical and Earth Sciences, University of Glasgow, Glasgow, U.K.
17
18

19
20
21 **ABSTRACT**

- 22
23 1. The complex life cycle of the globally threatened *M.margaritifera* includes a
24 parasitic stage, where glochidia attach to the gills of fishes of the *Salmo* genus.
25
26 However the species utilised appears to vary across its range. In previous literature,
27 the reported primary host in Scotland, home to a high proportion of the world's
28 remaining *M.margaritifera* populations, is the Atlantic salmon *Salmo salar* and in
29 its absence, the brown trout *Salmo trutta*.
30
31 2. In this study, the prevalence of infection in putative *Salmo* hosts in eight rivers in
32 NW Scotland was determined. At a selected site on each river, where both *S. trutta*
33 and *S. salar* were collected in abundance, *S. trutta* was the preferred host.
34
35 3. However at sites where *S. salar* were abundant but *S. trutta* were at low density, *S.*
36 *salar* showed a high prevalence of infection (with the exception of one river where
37 neither *S.salar* or *S.trutta* were infected). Thus the primary host appears to be very
38 site specific in the rivers sampled.
39
40 4. We speculate that this may be because *M. margaritifera* have population specific
41 responses to cues for attachment to a host. Alternatively it may be that host
42 population specific immune responses mediate infections by glochidia.
43
44 Additionally, larger fish were less likely to be infected than smaller fish and gills 1
45 and 5 were less infected than gills 2 to 4.
46
47 5. One consequence of this finding for both national and international conservation
48 management of this globally endangered species, is that any current or future
49
50
51
52
53
54
55
56
57
58
59
60

1
2
3 30 management activity must take into account local population host preferences,
4 31 otherwise conservation efforts may be in vain.
5
6 32

7
8 33 **Key words: Atlantic salmon, brown trout, freshwater pearl mussel, glochidia, host**
9 34 **fish, *Margaritifera margaritifera*, parasite, Scotland**
10

11 35
12 36 ***Correspondence to:** R. THOMAS, School of Geographical and Earth Sciences,
13 37 University of Glasgow, East Quadrangle, University Avenue, Glasgow, G12 8QQ. E-mail:
14 38 Rhian.Thomas@glasgow.ac.uk.
15
16 39

19 40 1. INTRODUCTION

20 41 The Freshwater Pearl Mussel, *Margaritifera margaritifera*, is a very long lived Unionid
21 42 Bivalve that is endangered and, as a result, highly protected across its range (IUCN, 2017;
22 43 Machordom, Araujo, Erpenbeck & Ramos, 2003; Ziuganov et al., 2000). Although
23 44 Scotland is a stronghold for this species, populations here as elsewhere, are showing
24 45 evidence of decline (Cosgrove et al., 2016; Skinner, Young & Hastie, 2003). A number of
25 46 causative factors have been suggested, including habitat degradation, pollution, and pearl
26 47 fishing. However the abundance of fish in rivers that also support *M.margaritifera* may
27 48 also be a factor in their decline (Hastie & Cosgrove 2001; Langan et al., 2007; Sime 2015).
28 49 For a very short but critical period in its early life cycle, *M.margaritifera* is parasitic.
29 50 Mature females release glochidia into the stream flow in summer; these are then carried in
30 51 water currents and taken into the gill chamber of host fish where they attach by clamping
31 52 their valves on to epithelial tissue of the gills (Meyers & Millemann, 1977).
32
33

34 53 This parasitic stage of the life cycle is not well understood. The host fish for
35 54 *M.margaritifera* are fish of the genus *Salmo*, however the species utilised appears to vary
36 55 across its range. Atlantic salmon (*Salmo salar*) is reported as the principal host of
37 56 *M.margaritifera* in Nova Scotia and Russia (Bauer, 1987). In contrast, brown trout (*Salmo*
38 57 *trutta*) appears to be the main host species in Germany and central Europe (Bauer, 1987).
39 58 There are only two *Salmo* species in Scotland, *S. salar* and *S.trutta* (Maitland & Campbell
40 59 1992). It has been reported that *S. salar* is the primary host but that *S. trutta* might act as a
41 60 sub-optimal host where the former species is absent for Scottish populations of
42 61 *M.margaritifera* (Hastie & Young 2001; Young & Williams 1984a; 1984b).The relative
43 62 importance of these two species as a host for *M.margaritifera*, is of considerable
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

1
2
3 63 management importance. The two species occupy different stream habitat types and there
4 64 is thus considerable variation in the relative abundance of each species at different rivers
5 65 and different sites within the same river (Klemetsen et al., 2003). This thus affects the
6 66 probability of contact between glochidia and a suitable host.

7
8
9 67 This study addressed four questions related to host utilisation by *M.margaritifera*:

- 10
11 68 1) Which of the two salmonid species is the preferred host for the parasitic stage of
12 69 the life cycle of the freshwater pearl mussels in selected rivers in north-west
13 70 Scotland?
14
15 71 2) Is host specificity consistent between *M. margaritifera* populations?
16
17 72 3) Is there a relationship between salmonid size and levels of glochidia encystment?
18
19 73 4) What component parts of the gill structure are parasitized?
20
21 74

22 75 2. METHODS

23
24 76 Potential fish hosts were collected from sites on eight rivers in north-west Scotland by
25 77 electrofishing between the 7th May 2013 and 20th June 2013. Osterling (2011) using a
26 78 similar method to that used here, indicated that June, the period immediately before
27 79 glochidia are shed (Hastie & Young, 2001) is the most appropriate period for visual
28 80 counting of glochidia. Similarly, Reid et al., (2013) conducted visual counting of glochidia
29 81 on salmonids in June. To protect *M.margaritifera* locations from potential illegal pearl-
30 82 harvesting, rivers are not named here but referred to only as rivers A-H. A suitable site on
31 83 each river was selected for its suitability as salmon and trout habitat and which were
32 84 located downstream of, and in close proximity (<30m) to *M.margaritifera* beds. Salmonid
33 85 fish were collected using a standard 500W DC backpack electro-fisher with one operator
34 86 and an assistant. Electrofishing has previously been shown to not adversely affect the
35 87 short-term survival of *M.margaritifera* (Hastie & Boon, 2001). Collected fish were
36 88 anaesthetised, identified, measured (fork length in mm) and the number of encysted
37 89 glochidia counted. At this time (immediately prior to excystment), glochidia were large
38 90 enough to count by eye. The fish were held in the hand on their dorsal surface and the
39 91 operculum gently lifted to make the gill filaments visible. Using a blunt needle to part the
40 92 gills it was possible to count individually encysted glochidia on the anterior and posterior
41 93 surfaces of the gill filaments separately for all five gills (from gill 1, to gill 5 (numbered
42 94 from anterior to posterior) on both left and right sides of the fish. Two people replicated
43 95 counts on a random sample of fish, to ensure accuracy and consistency of this visual count.
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

1
2
3 96 All fish were returned to the site on each river from which they were taken after a period of
4 97 recovery.

5
6 98 Glochidia frequencies were analysed using chi-squared analysis to test for
7
8 99 difference between host species and rivers. The total number of glochidia counted per fish
9
10 100 and the number on each gill arch was modelled in a general linear model (GLM) using fork
11 101 length and gill arch number as explanatory variables in R (Crawley, 2007). The relevance of
12 102 inclusion of each explanatory variable in the model was assessed in sequence using
13
14 103 significance testing between models (ANOVA; likelihood ratio tests [LRT]).
15

16 104 17 105 **3. RESULTS**

18
19 106 Across the eight rivers, a total of 830 fish (combined *S. salar* and *S. trutta*) were examined
20 107 for the glochidia of *M. margaritifera* (Table 1). The combined-river mean fork length (mm)
21 108 for *S. trutta* was 102.2 mm (+/- 23.1 S.D.) and *S. salar* 90.9mm (+/- 13.7). The overall mean
22 109 prevalence of infection in both species across the eight rivers was 14.7% (122 fish). No
23 110 infected fish of either species were detected in River E. If River E is excluded from
24 111 analysis, of the 740 remaining fish, 16.5% were infected and 83.5% uninfected.

25
26 112 Across all seven rivers where infection was detected, the combined prevalence of infection
27 113 for *S. trutta* (31.6% of 234 examined) was higher than that for *S. salar* (9.5% of 506
28 114 examined) ($\chi^2=435$; $df=1$; $P<0.0001$).

29
30 115 To examine the question of host specificity, data from five rivers are informative.
31 116 At rivers B, F and H both *S. trutta* and *S. salar* were collected in numbers large enough to
32 117 test for host use differences. At the sampled site within each of these rivers the prevalence
33 118 of infection of *S. trutta* (ranging from 11-65%) was statistically significantly higher than
34 119 for *S. salar* (0% for all three sites) (Table 1). In contrast, two additional rivers show a high
35 120 prevalence of infection by *S. salar*. The sampled site at rivers C and D showed 62% and
36 121 30% infection prevalence in *S. salar* respectively but there was only one *S. trutta* collected
37 122 (at river C) (Table 1), indicating that a very high prevalence of *S. salar* infection by
38 123 glochidia is possible at least when *S. trutta* are not available. At one further river (G)
39 124 infection prevalence of *S. trutta* was high (26%) but only four *S. salar* were collected so
40 125 there were no significant differences between species.

41
42 126 For *S. trutta* and *S. salar* combined, the number of glochidia detected on fish with
43 127 any infection, was significantly negatively related to fish fork length ($p<0.001$). Thus
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

1
2
3 128 smaller fish had significantly heavier glochidia load compared with fish with longer fork
4 129 lengths (Figure 1).

5
6 130 The GLM investigated the number of encysted glochidia on all five gills, and
7
8 131 revealed a significant two-way interaction between the side of the gill filament that was
9
10 132 infected (anterior or posterior) and the gill number (one to five)(Table2), but left/right side
11 133 of the fish was not significant. A *post hoc* Tukey test revealed there to be significantly
12 134 more encysted glochidia on gills two, three and four (which did not differ from each other)
13
14 135 than on gills one and five. In addition, *post hoc* testing also showed that there were
15
16 136 significantly more encysted glochidia on the anterior surface of gills two, three and four (p
17 137 <0.001 , $p<0.001$ and $p<0.01$) compared with the posterior side (Table 2).

18
19 138

20
21 139

4. DISCUSSION

22 140 In Scotland, literature indicates that *S. salar* is the primary host for *M. margaritifera*
23 141 glochidia, but where this species is not present, then *S. trutta* might be a sub-optimal host
24 142 (Hastie & Young, 2001; Young & Williams, 1984b). Data from the study presented here
25 143 shows that, not only is *S. trutta* a suitable host for the glochidia stage of the life cycle but
26
27 144 that, at least at some sampled sites within rivers, *S. trutta* is the preferred host even when
28
29 145 there are potential *S. salar* hosts present. Although only seven rivers with glochidia
30 146 infection were examined here, at two of these *S. salar* was the main host infected. At both
31
32 147 of these however, *S. trutta* density was too low to determine if *S. trutta* would have been
33
34 148 the principal host species if present and thus if *S. salar* was the optimal or a sub-optimal
35 149 host. Thus one conclusion of the study presented here is that there appears to be one
36
37 150 dominant host species for the parasitic phase of the *M. margaritifera* life-cycle, but that the
38 151 primary host used varies between each river. Thus at any one river, *salar* or *S. trutta* will
39
40 152 be the primary host and carry the bulk of the infections. Similar findings have been
41
42 153 reported from Scandinavia (Karlsson, Larsen, & Hindar, 2014; Larsen, Hårsaker, Bakken
43 154 & Barstad (2000); Salonen, Luhta, Moilanen, Oulasvirta, Turunen & Taskinen, 2017).
44
45 155 Whilst effort was made to ensure that selected sites on each river had both salmon and
46
47 156 trout juvenile habitat, it is acknowledged that the relative abundance of each species at
48
49 157 each site is likely to reflect local habitat type. However for five rivers sufficient numbers of
50
51 158 both species were collected to make meaningful between species comparison. The level of
52
53 159 glochidia encystment at any site is also likely to be affected by the proximity of
54
55 160 *M.margaritifera* beds to suitable juvenile fish habitat. Previous work showed the closer

1
2
3 161 mussel beds and good juvenile fish habitat were, the higher the levels of encystment
4 162 (Cosgrove & Hastie, 2001; Hastie, Watt & Cosgrove, 2011). In order to minimise the
5 163 capture of fish that were not infected efforts were thus made to electrofish within 30m or
6 164 less downstream of mussel beds to capture as high glochidia encystment as possible. The
7 165 timing of glochidia counts may also have an effect. It is possible that some glochidia may
8 166 have already dropped of the fish by June, however previous studies indicate that this period
9 167 is the time when such counts may be made most successfully (Osterling, 2011). These
10 168 caveats to this study may well have affected the absolute count number on fish presented
11 169 here but they are highly unlikely to affect the between species differences in infection
12 170 unless it is argued that the timing of excystment differs between species.

13
14
15
16
17
18
19 171 The mechanism through which glochidia may infect only, or almost only, one
20 172 *Salmo* species when two are present is uncertain. The infection process is thought to be
21 173 largely passive, in that glochidia are taken into the gill cavity and exposed to the gill
22 174 epithelium through the normal respiration process in the host fish (Meyers & Millemann
23 175 1977). However, there are at least two possible routes through which the observed
24 176 selectivity may occur. The glochidia may fail to attach to the exposed epithelium at the
25 177 appropriate time, if exposed to the “wrong” *Salmo* host. Alternatively the glochidia may
26 178 attach to its host but the host may initiate an immune response causing the shedding of
27 179 attached glochidia.

28
29
30
31
32
33 180 There is some circumstantial evidence in support of the first of these possibilities.
34 181 Karlsson and colleagues (Karlsson et al. 2014) examined the population genetics of
35 182 mussels from rivers where the prevalence of either *S. trutta* infection or *S. salar* infection
36 183 dominated. They concluded that *M. margaritifera* from *S. trutta* dominated host
37 184 populations were genetically different from populations which were *S. salar* host
38 185 dominated. One potential inference from this study is that there are significant genetic
39 186 differences between populations of *M. margaritifera* and that populations respond
40 187 differently to each of the putative *Salmo* hosts. In the UK a study (Cauwelier, Verspoor,
41 188 Tarr, Thomson, & Young, 2009) has shown that freshwater pearl mussel populations show
42 189 a major evolutionary split into northern and southern phylogenetic groups. They also found
43 190 that mussels from different river systems belonged to separate breeding populations and
44 191 the levels of genetic diversity within breeding populations varied significantly and were
45 192 higher in Scotland, compared to England and Wales.

1
2
3 193 There is also similar circumstantial support for the second of these two
4 194 possibilities. *S. salar* is known to exhibit significant between-population variation in genes
5 195 of the Major Histocompatibility Complex, which has consequences for immune-
6 196 competence. This is assumed to be a response to the differential exposure of populations to
7 197 parasites (Dionne, Miller, Dodson, Caron & Bernatchez, 2007). It is a reasonable
8 198 proposition to suggest that discrete *Salmo* populations may similarly show different
9 199 response capabilities on exposure to *M. margaritifera* glochidia as a result of their separate
10 200 evolutionary past. In addition, the study presented here showed that larger and thus older
11 201 fish had lower infection levels than smaller and younger fish. Although speculative, one
12 202 possibility is that older fish acquire some immunity as a result of previous exposure that
13 203 offers some protection from subsequent glochida infection, as has been shown in a
14 204 previous study (Thomas, Taylor & Garcia de Leaniz, 2014).

15 205 Data presented here also shows that *M. margaritifera* infection does not affect all
16 206 gills equally. The outer gill arch and the inner gill arch had lower levels of infection than
17 207 the middle three arches. It is likely that the inner arch is partly protected by its relative
18 208 position behind the four others and that the first gill arch might be less infected as a result
19 209 of the abrasion effect of the opercular bone on glochidia.

20 210 There are a number of important management consequences that stem from these
21 211 findings. Firstly because the host species may not be the same for each population, then
22 212 any assessment of the density of hosts as one possible stressor on a population of *M.*
23 213 *margaritifera* has to take account of the *Salmo* species that is the relevant host for that
24 214 specific *M. margaritifera* population. It cannot be assumed that one host species is a simple
25 215 suitable replacement for the other. Secondly, the very clear host specificity for one or the
26 216 other of the two *Salmo* species reported here and elsewhere (see Bauer, 1987) suggests that
27 217 if *M. margaritifera* glochidia are forced to use the less preferred host species at any river
28 218 (because density of the preferred host is very low) then there may be some consequences
29 219 for individual survival of animals in the longer term.

30 220

31 221

5. ACKNOWLEDGEMENTS

32 222 Funding was provided by the European Union INTERREG IVA Programme (project 2859
33 223 'IBIS') managed by the Special EU Programmes Body and the Scottish Government
34 224 project SP004. We thank Scottish Natural Heritage, District Salmon Fisheries Boards,

1
2
3 225 proprietors and landowners for access, and Bruce Wallace and Jennifer Dodd for assistance
4 226 with data collection.
5
6

7 227
8

9 228

6. REFERENCES

10 229 Bauer, G. 1987. The parasitic stage of the freshwater pearl mussel (*Margaritifera*
11 230 *margaritifera* L.) 3 Host relationships. *Archiv für Hydrobiologie*, 76, 413–423.

12
13
14 231 Cawelier, E., Verspoor, E., Tarr, E.C., Thomson, C & Young, M. 2009. Genetic diversity
15 232 and differentiation of the freshwater pearl mussel (*Margaritifera margaritifera*)
16 233 populations in the UK. Scottish Natural Heritage Commissioned Report No 344
17 234 (ROAME No FO5AC701).

18
19
20 235 Cosgrove, P. & Hastie, L.C., 2001. Conservation of threatened freshwater pearl mussel
21 236 populations: river management, mussel translocation and conflict resolution.
22 237 *Biological Conservation*, 99, 183-190.

23
24
25 238 Cosgrove, P., Watt, J., Hastie, I., Sime, I., Shields, D., Cosgrove, C., Brown, L.,
26 239 Isherwood, I., & Bao, M. 2016. The status of the freshwater pearl mussel
27 240 *Margaritifera margaritifera* in Scotland: extent of change since 1990s, threats and
28 241 management implications. *Biodiversity and Conservation*, 25, 2093-2112.

29
30 242 Crawley, M.J. 2007. The R Book. John Wiley and Sons, Chichester.

31
32 243 Dionne, M., Miller, K.M., Dodson, J., Caron, F., Bernatchez, L. 2007. Clinal variation in
33 244 MHC diversity with temperature: evidence for the role of host–pathogen interaction
34 245 on local adaptation in Atlantic salmon. *Evolution*, 61, 2154-2164.

35
36 246 Hastie, L.C. & Boon, P. 2001. Does electrofishing harm freshwater pearl mussels? *Aquatic*
37 247 *Conservation: Marine and Freshwater Ecosystems*, 11, 149-152.

38
39 248 Hastie, L.C., & Cosgrove, P. 2001. The decline of migratory salmonid stocks: A new threat
40 249 to pearl mussels in Scotland. *Freshwater Forum*, 15, 85–96.

41
42 250 Hastie, L.C., & Young, M.R. 2001. Freshwater pearl mussel (*Margaritifera margaritifera*)
43 251 glochidiosis in wild and farmed salmonid stocks in Scotland. *Hydrobiologia*, 445, 109–
44 252 119.

45
46 253 Hastie, L.C., Watt, J. & Cosgrove, P.J., 2011. Restoration of freshwater pearl mussel in
47 254 selected Scottish rivers: phase 2b – factors determining the success of restoration
48 255 measures. Scottish Natural Heritage Commissioned Report No.458.
49
50
51
52
53
54
55
56
57
58
59
60

- 1
2
3 256 IUCN 2017. IUCN Red List of Threatened Species. Version 2017.1. www.iucnredlist.org.
4 257 Accessed 26 May 2017.
5
6 258 Karlsson, S., Larsen, B.M., & Hindar, K. 2014. Host-dependent genetic variation in
7 259 freshwater pearl mussel (*Margaritifera margaritifera* L.). *Hydrobiologia*, 735, 179–
8 260 190.
9
10
11 261 Klemetsen, A., Amundsen, P.A., Dempson, J.B., Jonsson, B. Jonsson, N., O’Connell,
12 262 M.F. & Mortensen, E. 2003. Atlantic salmon (*Salmo salar* L.), brown trout (*Salmo*
13 263 *trutta* L.) and Arctic charr (*Salvelinus alpinus* L.): a review of aspects of their life
14 264 histories. *Ecology of Freshwater Fish*, 12, 1–59.
15
16
17 265 Langan, S., Cooksley, S., Young, M., Stutter, M., Scougall, F., Dalziel, A., Feeney, I.,
18 266 Lilly, A. & Dunn, S. 2007. The management and conservation of the freshwater pearl
19 267 mussel in Scottish catchments designated as Special Areas of Conservation or Sites of
20 268 Special Scientific Interest. Scottish Natural Heritage Commissioned Report No.249
21 269 (ROAME No. F05AC607).
22
23
24 270 Larsen, B. M., Hårsaker, K., Bakken, J. & Barstad, D.V. 2000. The freshwater pearl
25 271 mussel *Margaritifera margaritifera* in Steinkjervassdraget and Figga, Nord-Trøndelag.
26 272 Preliminary survey in connection with planned rotenone treatment. NINA Fagrapport
27 273 39: 1–39. [In Norwegian with English summary].
28
29
30 31
32 274 Maitland, P.S., & Campbell, R.N. 1992. Freshwater Fishes. London: Harper Collins.
33
34 275 Machordom, A., Araujo, R., Erpenbeck, D. & Ramos, M.A. 2003. Phylogeography and
35 276 conservation genetics of endangered European Margaritiferidae (Bivalvia:
36 277 Unionoidea). *Biological Journal of the Linnean Society*, 78, 235–252.
37
38 278 Meyers, T.R. & Millemann, R.E. 1977. Glochidiosis of salmonid fishes. I. Comparative
39 279 susceptibility to experimental infection with *Margaritifera margaritifera* (L.)
40 280 (Pelecypoda: Margaritanidae). *Journal of Parasitology*, 63, 728–733.
41
42
43 281 Österling, M.E. 2011. Test and application of a non-destructive photo-method investigating
44 282 the parasitic stage of the threatened mussel *Margaritifera margaritifera* on its host fish
45 283 *Salmo trutta*. *Biological Conservation*, 144, 2984–2990.
46
47
48 284 Reid, N., Keys, A., Preston, J.S., Moorkens, E., Roberts, D. & Wilson, C.D. 2013.
49 285 Conservation status and reproduction of the critically endangered freshwater pearl
50 286 mussel (*Margaritifera margaritifera*) in Northern Ireland. *Aquatic Conservation:*
51 287 *Marine and Freshwater Ecosystems*, 23, 571–581.
52
53
54
55
56
57
58
59
60

- 1
2
3 288 Salonen, J., Luhta, P-L., Moilanen, E., Oulasvirta, P., Turunen, J. & Taskinen, J. 2017.
4 289 Atlantic salmon (*Salmo salar*) and brown trout (*Salmo trutta*) differ in their suitability
5 290 as hosts for the endangered freshwater pearl mussel (*Margaritifera margaritifera*) in
6 291 northern Fennoscandian rivers. *Freshwater Biology*, 62, 8, 1346-1358.
- 9 292 Sime, I. 2015. Freshwater pearl mussel. Version 1.0. In The Species Action Framework
10 293 Handbook, Gaywood MJ, Boon PJ, Thompson DBA, Strachan IM (eds). Scottish
11 294 Natural Heritage, Battleby, Perth.
- 14 295 Skinner, A., Young, M.R., & Hastie, L.C. 2003. Ecology of the Freshwater Pearl Mussel.
15 296 Conserving Natura 2000 Rivers Ecology Series No.2 English Nature, Peterborough.
- 17 297 Thomas, G.R., Taylor, J., & Garcia de Leaniz, C. 2014. Does the parasitic freshwater pearl
18 298 mussel *M. margaritifera* harm its host? *Hydrobiologia*, 735, 191–201.
- 20 299 Young, M.R. & Williams, J. 1984a. The reproductive biology of the freshwater pearl
21 300 mussel *Margaritifera margaritifera* in Scotland 2. Laboratory Studies. *Archiv für*
22 301 *Hydrobiologie*, 100, 29–43.
- 25 302 Young, M.R. & Williams, J. 1984b. The reproductive biology of the freshwater pearl
26 303 mussel *Margaritifera margaritifera* in Scotland. 1. Field Studies. *Archiv für*
27 304 *Hydrobiologie*, 99, 405–422.
- 30 305 Ziuganov, V., Miguel, E.S., Neves, R.J., Longa, A., Fernández, C., Amaro, R., Beletsky,
31 306 V., Popkovitch, E., Kaliuzhin, S., & Johnson, T. 2000. Life Span Variation of the
32 307 Freshwater Pearl Shell: A Model Species for Testing Longevity Mechanisms in
33 308 Animals. *AMBIO: A Journal of the Human Environment*, 29, 102.

36 309
37 310
38 311
39 312
40 313
41 314
42 315
43 316
44 317
45 318
46 319
47 320
48
49
50
51
52
53
54
55

321 Table 1. The prevalence of infection and sample size in two *Salmo* species in eight rivers
 322 in north-west Scotland.

River	<i>S. trutta</i>	<i>S. trutta</i> N	<i>S. trutta</i> %	<i>S. salar</i>	<i>S. salar</i> N	<i>S. salar</i> %
	N	Infected	Infected	N	Infected	Infected
A	40	22	55	2	0	0
B	23	15	65	232	0	0
C	1	0	0	55	34	62
D	0	0	0	46	14	30
E	0	0	0	90	0	0
F	21	4	19	122	0	0
G	113	29	26	4	0	0
H	36	4	11	45	0	0

323

324

325

326

327 Table 2. Summary of the total numbers of encysted glochidia counted on anterior and
 328 posterior sides of the gill filaments on each of the five gills of infected *S. trutta* and *S.*
 329 *salar*.

330

Gill Number	Mean glochidia count		Standard deviation	
	Anterior	Posterior	Anterior	Posterior
1	0.0	0.0	0.0	0.0
2	11.3	3.7	15.1	6.9
3	10.8	5.0	14.5	8.4
4	9.8	5.2	13.1	7.9
5	6.4	3.5	8.7	6.4

331

332

333

1
2
3 334

4 335 Figure 1. Relationship between the fork length (mm) of individual fish caught and the total
5 336 glochidia counted ($F=96.43$, $df=1$, $r^2 = 0.06$, $p < 0.001$).
6 337

7
8 338 (Separate jpeg file)
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

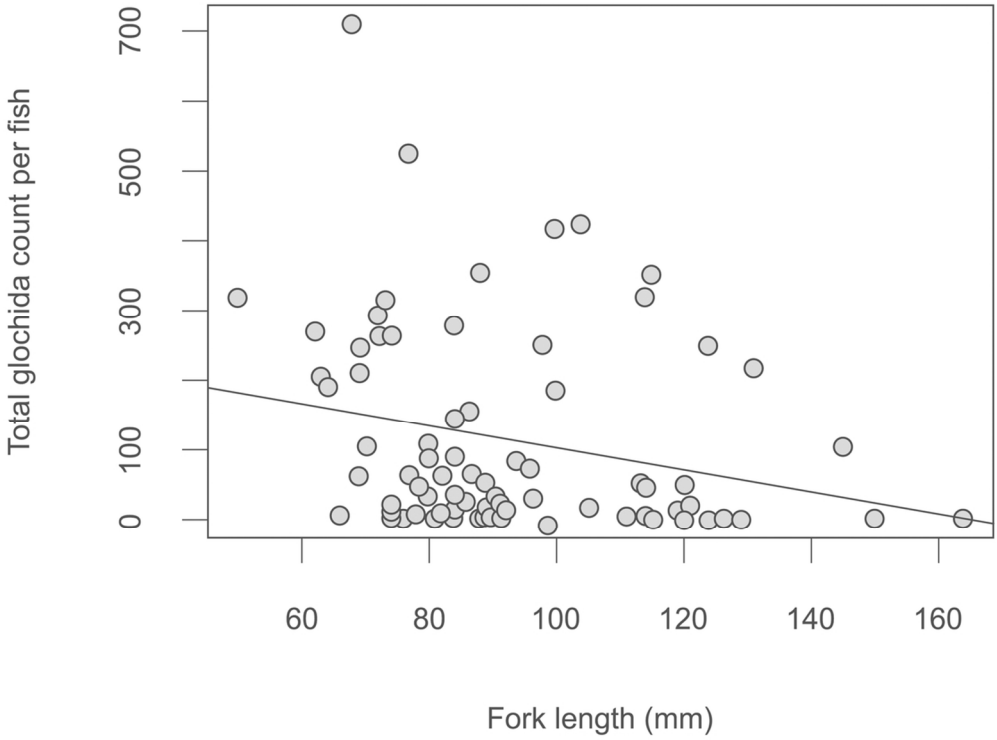


Figure 1. Relationship between the fork length (mm) of individual fish caught and the total glochidia counted (F=96.43, df =1, r 2 = 0.06, p =<0.001).

107x79mm (300 x 300 DPI)