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1 **Seaweed fertilisation impacts the chemical and isotopic**
2 **composition of barley: Implications for analyses of archaeological**
3 **skeletal remains**

4 Magdalena Blanz^{a,b,*}, Philippa Ascough^c, Ingrid Mainland^a, Peter Martin^d, Mark A. Taggart^e,
5 Burkart Dieterich^{d,f}, John Wishart^d, Kerry L. Sayle^c, Andrea Raab^b, Jörg Feldmann^b

6

7 ^aArchaeology Institute, University of the Highlands and Islands, Orkney College UHI, East
8 Road, Kirkwall, Orkney, KW15 1LX, Scotland, UK

9 ^bTrace Element Speciation Laboratory (TESLA), Department of Chemistry, University of
10 Aberdeen, Meston Building, Meston Walk, Aberdeen, AB24 3UE, Scotland, UK

11 ^cScottish Universities Environmental Research Centre, Rankine Avenue, Scottish Enterprise
12 Technology Park, Rankine Av, East Kilbride G75 0QF, UK

13 ^dAgronomy Institute, University of the Highlands and Islands, Orkney College UHI, East Road,
14 Kirkwall, Orkney, KW15 1LX

15 ^eEnvironmental Research Institute, University of the Highlands and Islands, Castle Street,
16 Thurso, Scotland, KW17 7JD, UK

17 ^fThe James Hutton Institute, Errol Road, Dundee, DD2 5DA, UK

18

19 *Corresponding author: m.blanz.11@aberdeen.ac.uk

20 **Address:** Archaeology Institute
21 University of the Highlands and Islands
22 Orkney College UHI
23 Kirkwall, Orkney
24 KW15 1LX
25 Scotland, UK

26 **Abstract**

27 Fertilisation with animal manure has been shown to affect crop chemical and isotopic
28 composition, indicating that if manuring effects are not taken into account, there is a risk of
29 overestimating consumer trophic levels in palaeodietary studies. The effect of fertilisation
30 with seaweed, a common fertiliser in the past in coastal areas, has been the subject of
31 several hypotheses, but until now has not been studied in this particular context.

32 In this study, the impact of fertilising *bere*, an ancient type of Scottish barley (*Hordeum*
33 *vulgare* L.), with 25 t/ha and 50 t/ha seaweed, in comparison to a modern commercial
34 mineral fertiliser and to no fertilisation, was investigated in a field trial on the Orkney
35 Islands, Scotland. Stable isotope ratios ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) and elemental compositions (B, Mg,
36 K, Ca, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, As, Sr, Mo, Cd and Pb) of grain, husk and straw samples
37 were determined. Significant differences were found between treatment groups, including
38 increases in $\delta^{15}\text{N}$ values of $0.6 \pm 0.5 \text{ ‰}$ (average $\pm 1\sigma$ for five replicate plots) in grain, and
39 $1.1 \pm 0.4 \text{ ‰}$ in straw due to seaweed fertilisation. Elevated concentrations of Sr in grain and
40 husk samples (factors of 1.2 to 1.4) indicate the geographic tracer $^{87}\text{Sr}/^{86}\text{Sr}$ may also be
41 affected.

42 Fertilisation with seaweed thus needs to be considered for archaeological interpretations of
43 chemical and isotopic compositions of crop and skeletal material for accurate palaeodietary
44 and provenance reconstructions, particularly in coastal areas. Further implications of these
45 results for studies concerning the effects of sea spray, radiocarbon-dating, and for dietary
46 reconstructions using trace elements are also identified.

47

48 **Keywords:**

49 manuring

50 kelp fertiliser

51 coastal archaeology

52 past/prehistoric agriculture

53 crop husbandry

54 land management

55 archaeological chemistry

56 1 Introduction

57 The study of archaeological skeletal material using stable isotope ratio and trace elemental
58 analysis has frequently been used to infer past diets and geographic origin of humans and
59 animals (reviewed in e.g. Bentley, 2006; Lee-Thorp, 2008). These dietary reconstructions are
60 based on the predictable transfer of a chemical or isotopic "signature" from the diet to the
61 skeleton during life. However, for such research to be robust, it is necessary to have a
62 thorough understanding of how the chemical and isotopic composition of skeletal material
63 is influenced by naturally (e.g. climate, underlying geology; Bentley, 2006; Craine et al.,
64 2009) and anthropogenically (e.g. fertilisation, irrigation; Bogaard et al., 2007) induced
65 variability in the composition of primary producers such as cereals, trees and even algae.
66 Understanding the extent and origin of such variability and how it is transferred up the food
67 chain greatly improves the accuracy of dietary reconstructions of humans and animals
68 (Tieszen, 1991; van Klinken et al., 2000).

69 The importance of taking manuring in particular into account is well-illustrated when
70 considering nitrogen stable isotope ratios ($\delta^{15}\text{N}$), which are commonly used as indicators of
71 trophic level as they reflect $\delta^{15}\text{N}$ of dietary protein, but additionally increase up the food
72 chain by generally around 3–5 ‰ per trophic level in skeletal collagen (Bocherens and
73 Drucker, 2003; Hedges and Reynard, 2007). Fertilisation with animal dung has been shown
74 to elevate crop $\delta^{15}\text{N}$ values by up to (or potentially more than) 7 ‰ compared to
75 unfertilised crops (Bogaard et al., 2007; Bol et al., 2005; Commisso and Nelson, 2007; Fraser
76 et al., 2011; Kanstrup et al., 2012, 2011; Styring et al., 2014a; Treasure et al., 2016). This
77 leads to elevated $\delta^{15}\text{N}$ values in consumers (particularly when plants are the dominant
78 protein source). Additionally, after consumption by e.g. sheep, this elevation in $\delta^{15}\text{N}$ values
79 can be passed up the food chain in the form of dietary protein. Thus, when manuring is not
80 taken into account, there is a danger of overestimating the trophic levels of all consumers in
81 the food chain, including those who do not directly consume fertilised plants in substantial
82 amounts but do consume animal products.

83 Fertilisation with seaweed can significantly increase yields of various terrestrial crops (Khan
84 et al., 2009), and its historic use as a fertiliser has been documented in Europe (e.g. Arzel,
85 1984; Kenicer et al., 2000; Russell, 1910), Asia (e.g. Komatsu and Yanagi, 2015; Maddison,
86 2006; Tajima, 2007) and America (e.g. Mikkelsen and Bruulsema, 2005; Suttles, 2005;
87 Thompson, 2005). Widely available on rocky shores, seaweed would have been especially
88 valuable in the past in areas where the amount of livestock kept could not provide sufficient
89 dung. Utilising seaweed instead of dung as fertiliser also relaxed constraints on livestock
90 management, e.g. allowing for the out-wintering of stock, as seaweed obviated the need to
91 collect dung by housing animals over winter (Dodgshon, 2011; Zimmermann, 1998).
92 Additionally, seaweed has also been reported to be preferable to dung as a fertiliser
93 because seaweed does not tend to harbour pathogens harmful to terrestrial plants, or
94 introduce weeds via undigested seeds (Hendrick, 1898).

95 While numerous modern agronomic studies have investigated the use of seaweed as
96 fertiliser (reviewed in Khan et al., 2009), past marine plant use is not currently widely

97 researched (but this is beginning to change; e.g. [Mooney, 2018](#)) and archaeologically
98 important effects on crop composition (particularly $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$) have not yet been
99 studied. Stable carbon isotope ratios ($\delta^{13}\text{C}$) are often used to distinguish between terrestrial
100 and marine foods, since in absence of C_4 plants, collagen $\delta^{13}\text{C}$ values of -12‰ generally
101 indicate almost all dietary protein to be marine, while values of -20‰ indicate diets
102 without significant amounts of marine protein (Richards and Hedges, 1999). It has been
103 suggested that fertilisation with marine products (particularly seaweed) may lead to
104 elevated crop $\delta^{13}\text{C}$ values (Craig et al., 2005; Jones and Mulville, 2016; Milner et al., 2004;
105 Murray et al., 2012), which, if unaccounted for, would lead to an overestimation of the
106 direct consumption of marine foods. However, as terrestrial plants primarily acquire carbon
107 by photosynthesis with atmospheric CO_2 , rather than from soil, it has also been asserted
108 that fertilisation with marine material does not affect crop $\delta^{13}\text{C}$ values (Fraser et al., 2017;
109 Richards and Schulting, 2006; Schulting et al., 2010).

110 Other hypothesised effects concerning marine-fertilised terrestrial crops include increased
111 $\delta^{15}\text{N}$ values (Fraser et al., 2017; Jones and Mulville, 2016; Schulting and Richards, 2009;
112 Schulting et al., 2010), increased $\delta^{34}\text{S}$ values (Fraser et al., 2017; Lamb et al., 2012; Schmidt
113 et al., 2005), increased strontium (Sr) concentrations and a shift toward marine $^{87}\text{Sr}/^{86}\text{Sr}$
114 isotope ratios (Evans et al., 2012; Montgomery et al., 2007, 2003; Montgomery and Evans,
115 2006).

116 Clarity as to the effects of seaweed fertilisation on the chemical and isotopic composition of
117 terrestrial crops would aid in the interpretation of existing and future isotope ratio and trace
118 elemental data. This could contribute to e.g. the European Neolithic–Mesolithic transition
119 debate, wherein the dietary importance of marine resources in particular has long been
120 discussed: It has been argued that marine resources were important in the Mesolithic, but
121 abruptly lost significance once farming began in the Neolithic (e.g. Cramp et al., 2014;
122 Richards and Schulting, 2006; Schulting and Richards, 2002). This has also been interpreted
123 to imply a type of taboo surrounding marine foods in the Neolithic (Thomas, 2003). Others
124 have argued against this, reasoning that marine resources continued to be exploited in the
125 Neolithic in significant amounts (Lidén et al., 2004; Milner et al., 2006, 2004) and may have
126 been particularly important during famines in adverse climates (Montgomery et al., 2013).

127 However, in these discussions, the term “marine” is usually used to refer to marine
128 mammals, fish and shellfish, and seaweed has largely been ignored both as a source of food
129 for humans and animals, and as a fertiliser. This is likely in part due to the difficulty of
130 identifying contributions of seaweed to complex diets, both by isotopic measurements and
131 other means (though Neolithic Orkney sheep have recently been shown to have been
132 consuming seaweed; Balasse et al., 2009; Schulting et al., 2017). Thus, new approaches are
133 needed to identify seaweed consumption, which may include e.g. studies of the elemental
134 composition of tooth enamel in seaweed-eating vertebrates. Such approaches would
135 however require modern baseline data for marine, coastal and terrestrial ecosystems, as
136 well as data from seaweed-fertilised plants, if informed interpretations of archaeological
137 data are to be made.

138 In this study, our aim is to explore the effect of seaweed fertilisation on $\delta^{15}\text{N}$, $\delta^{13}\text{C}$ and
139 elemental composition of the crops by performing a field trial growing bere, a Scottish
140 barley (*Hordeum vulgare* L.) landrace, with seaweed fertilisation. This will establish modern
141 baseline data for marine-fertilised terrestrial crops, aiding in more accurate interpretations
142 of the chemical and isotopic compositions of skeletal remains of (potential) direct and
143 indirect consumers of such crops, as well as crop husbandry practices.

144 2 Historical and archaeological background to field trial design

145 The field trial was designed to be similar to historically documented seaweed fertilisation
146 practices, whilst taking practicability into account. Bere barley, a hulled lax-eared six-row
147 landrace of barley, was chosen as the crop for this field trial due to the particular
148 importance of barley for both human and animal consumption in Northern Europe from the
149 Neolithic onwards (Bishop et al., 2009; Dockrill et al., 1994; Hunter et al., 1993; McClatchie
150 et al., 2014). Bere barley is one of the oldest cereals still in cultivation in Britain (Jarman,
151 1996; Martin et al., 2008; Wallace et al., 2018) making it more likely to be similar to barley
152 found archaeologically than modern barley varieties. Numerous historical sources indicate
153 that barley was frequently fertilised with seaweed (e.g. Fenton, 1997; Martin, 1716; Russell,
154 1910; Sauvageau, 1920).

155 The choice of seaweed for fertilisation ranged widely, with local preferences for either cut
156 or stranded seaweed, and for specific species (e.g. *Laminaria spp.*, *Fucus spp.*, *Ascophyllum*
157 *nodosum*; Fenton, 1997; Hendrick, 1898; Neill, 1970; Russell, 1910; Sauvageau, 1920). Due
158 to the lack of consensus as to which species of seaweed is/was historically preferred for
159 fertilisation, and since all the preferred species are abundant on rocky shores in Britain and
160 Ireland today (Hardy and Guiry, 2003) and have likely been for the past 6,000 years (Coyer
161 et al., 2003; Muhlin and Brawley, 2009; Olsen et al., 2010; Rothman et al., 2017), we
162 decided for practical reasons to use stranded seaweed of various species (including e.g.
163 *Laminaria spp.*, *Fucus spp.*, *Ascophyllum nodosum*), as found on the shore, for this field trial.
164 Historical seaweed application rates documented in the literature ranged from 10 t/ac to 50
165 t/ac (ca. 25 t/ha to ca. 124 t/ha; Hendrick, 1898; Noble, 1975; Russell, 1910; Stephenson,
166 1968). The selected application rate presumably mainly depended on the availability of
167 labour, draught animals and seaweed, as well as the type and quality of the soil, and the
168 crop type. In the case of bere barley, over-fertilisation leads to increased incidences of
169 lodging (i.e. falling over), which can negatively impact plant growth and complicates
170 harvesting (Shah et al., 2017). Hence, two rather conservative application levels of 25 t/ha
171 and 50 t/ha seaweed (i.e. ca. 10 and 20 t/ac) were chosen for this field trial.

172 Historically, seaweed application was often undertaken multiple times a year, with seaweed
173 generally applied fresh from the shore in autumn or winter, and as compost when the crop
174 was about to be seeded or already growing (Dodgshon, 1988; Fenton, 1997; Noble, 1975;
175 Russell, 1910; Sauvageau, 1920; Stephenson, 1968). For this study, seaweed was composted
176 and applied shortly before sowing. A modern commercial fertiliser was also used in this

177 study on separate plots to help distinguish between the more general effects of fertilisation,
178 and effects that are specific to fertilisation with seaweed.

179 3 Materials and methods

180 3.1 Field trial design and implementation

181 An agronomic experimental site ca. 100 m north of Orkney College UHI (Scotland) and ca.
182 250 m south of the nearest coastline was chosen for the field trial (58° 59' N and 2° 57' W;
183 grid reference HY 456 114). This area has an acidic clay loam soil (see supplementary
184 material). In previous years, the field had been cultivated and fertilised with a NPK mineral
185 fertiliser at a low level of 50 kg N/ha (likely with a $\delta^{15}\text{N}$ value between 0 and -1 ‰, Bateman
186 and Kelly, 2007; described further below). No other fertilisation-based agronomic field trials
187 had been performed in this area before, so that the soil was considered largely
188 homogeneous throughout the trial area.

189 The trial plots were laid out in a randomised block design as 3 m × 3 m (9 m²) plots, with 1 m
190 space between adjacent plots and five replicate plots per fertilisation treatment. Around
191 450 kg of stranded seaweed of various species were collected from Newark Bay, Mainland,
192 Orkney (Grid reference: HY 567 041). After composting for 1.5 months in aerated plastic
193 bags, the composted seaweeds were manually evenly distributed onto marked out plots on
194 the ploughed, power-harrowed field at rates of 25 t seaweed/ha and 50 t seaweed/ha (wet
195 weight; corresponding to ca. 200 kg N/ha and 400 kg N/ha, not all of which was
196 bioavailable). A conventional 14-14-21 NPK fertiliser (YaraMila MAINCROP 14-14-21; Yara
197 UK Ltd, Belfast, UK) was manually applied to a third set of plots at 50 kg N/ha. A fourth set
198 of plots (control plots) were not fertilised in any way, making up a total of 20 plots (5
199 unfertilised, 5 with 25 t seaweed/ha, 5 with 50 t seaweed/ha, 5 NPK-fertilised). After
200 spreading the fertilisers, all plots were power-harrowed twice to mix the seaweeds into the
201 soil. The barley was sown the following day (early May 2017) at a rate of approximately 16
202 g/m² with a thousand grain weight of 30.3 g, using a tractor drawn seeder (width 3 m). The
203 soil surface was then flattened using a Cambridge roller. After one month of growth a
204 herbicide mixture (see supplementary material) was applied to all plots in order to prevent
205 excessive weed growth.

206 The bere barley was harvested in early September 2017 from a 1 m × 1 m square at the
207 centre of each 3 m x 3 m plot to avoid edge effects, issues related to soil compaction due to
208 tractor wheels, and effects due to fertiliser run-off. The harvested barley was dried at 30 °C
209 until constant weight (ca. 48 h) and weighed for yield evaluation. A random subsample of 15
210 stalks (including ears) was taken for chemical and isotopic analysis from each plot.

211 3.2 Chemical and isotopic analyses of bere barley

212 3.2.1 *Sample pre-treatment*

213 The harvested barley was separated into straw, grain (including bran) and husk samples for
214 analysis, as these different parts would have been consumed to different extents by humans
215 and livestock. From each of the 15 sampled ears per plot, all grains from half of the ear (top
216 to bottom) were manually separated from the rachis, and the awns were manually
217 separated from the husks. This resulted in samples of around 300 grains per plot, weighing
218 ca. 10 g per sample including the husk and bran. From this, a random subsample of
219 approximately 2 g of grain (ca. 50-70 grains) per plot was taken, from which husks were
220 manually removed and kept for analysis. As the bran was not easily removable and would
221 likely not (commonly) have been removed in the past (Britton and Huntley, 2011; Fenton,
222 1997; Jadhav et al., 1998), the de-husked grains were not treated further. Grains were then
223 homogenised by mortar and pestle. Around 10 g of dried straw from each plot was ground
224 using an electric spice and nut grinder (Model SG20U, Cuisinart Corp., Greenwich, USA), and
225 then sieved to 1 mm with a plastic mesh. This processing yielded five samples (one per
226 replicate plot) for each of the four treatment types per plant part, i.e. 20 unique samples for
227 each of husk, grain, and straw, all of which were analysed for their chemical and isotopic
228 composition as described below.

229 For the analysis of the fertilisers, a pooled sample (120 g dry weight) of the composted
230 seaweed as it was at the time of application in May was dried, ground and sieved as
231 described for the straw samples. An aliquot of 1.5 g sample of the conventional NPK
232 fertiliser was homogenised to a fine powder using a mortar and pestle.

233 3.2.2 *Elemental composition analysis*

234 The concentrations of B, Mg, K, Ca, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, As, Sr, Mo, Cd and Pb in
235 straw, grain, husk, and the seaweed and NPK fertilisers were determined. For this, 0.1 g of
236 each sample except the NPK fertiliser were left to pre-digest overnight with 2 mL HNO₃ (70
237 % analytical reagent grade, Fisher Scientific UK). After addition of 3 mL H₂O₂ (30 % w/v
238 laboratory reagent grade, Fisher Scientific UK), the samples were microwave digested using
239 a non-pressurized CEM Mars 5 system (CEM Microwave Technology Ltd., UK), with samples
240 heated to 95 °C for 30 min. Dilutions were then performed using bidistilled water (Aquatron
241 still A4000D, Bibby Scientific Limited, UK). The NPK fertiliser was prepared by addition of 13
242 mL bidistilled water and 1 mL concentrated HNO₃ to 0.1 g of sample, without microwave
243 digestion.

244 Analysis was performed by microwave plasma atomic emission spectroscopy (MP-AES;
245 Agilent 4200, instrument parameters in Table S.1, supplementary material) and by
246 inductively coupled plasma tandem mass spectrometry (ICP-MS/MS; Agilent 8800,
247 instrument parameters in Table S.2, supplementary material). Triplicate measurements
248 were performed every five samples. Certified reference materials NIST1568a (rice flour),
249 NIST1573a (tomato leaves), NIST3232 (kelp powder) and NIST8415 (whole egg powder),

250 which were microwave-digested and analysed as above, yielded recoveries of mainly
251 between 80 and 120 % (Tables S.3 and S.4, supplementary material).

252 3.2.3 *Stable isotope ratio analysis for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$*

253 The husk samples were comminuted using single edge razor blades (Fisher Scientific,
254 Loughborough, UK) on a granite cutting surface to a size where no spatial dimension was > 2
255 mm. Around 600 μg and 3–10 mg of each husk, grain, straw and seaweed sample (exact
256 weights known) were weighed into separate tin capsules for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ measurements,
257 respectively. Stable isotope ratios were determined using a Delta V Advantage continuous-
258 flow isotope ratio mass spectrometer coupled via a ConFlo IV to an IsoLink Elemental
259 Analyser (Thermo Scientific, Bremen). Triplicate measurements were performed every five
260 samples and after every ten unknown samples, in-house standards calibrated to the
261 international reference materials USGS40, USGS41, IAEA-CH-6 ($\delta^{13}\text{C}$ values -26.39‰ ,
262 $+37.63\text{‰}$, -10.45‰ , respectively), USGS25, IAEA-N-1 and IAEA-N-2 ($\delta^{15}\text{N}$ values -30.41‰ ,
263 $+0.43\text{‰}$, $+20.41\text{‰}$, respectively) were run in duplicate. Results are reported as permille (‰)
264 relative to the international reference standards VPDB and AIR with 1σ precisions of $\pm 0.2\text{‰}$
265 ($\delta^{13}\text{C}$) and $\pm 0.3\text{‰}$ ($\delta^{15}\text{N}$).

266 3.3 Data treatment

267 Analytical errors were calculated as 1σ of triplicate measurements of every fifth sample
268 analysed. To gain an overview of the data generated, principal component analysis (PCA;
269 Bro and Smilde, 2014; Wold et al., 1987) was performed based on a correlation matrix of the
270 determined elemental concentrations and stable isotope ratios ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) using
271 Minitab statistical software (Minitab 14, Minitab Inc., USA). Significant differences between
272 sample groups were assessed by one-way and two-way (fertilisation treatment and plant
273 part) ANOVA followed by post-hoc Tukey tests, as well as two-sample two-tailed t-tests
274 using Minitab. The statistical significance threshold was set at $\alpha = 0.05$.

275 4 Results

276 Fertilisation with all fertilisers led to an approximate doubling in the bere barley yield in
277 terms of both straw and ear weights per m^2 when compared to unfertilised plots.
278 Significantly higher ear weight per m^2 yields were observed for the 50 t/ha seaweed
279 treatment than the 25t/ha seaweed treatment (manuscript in preparation).

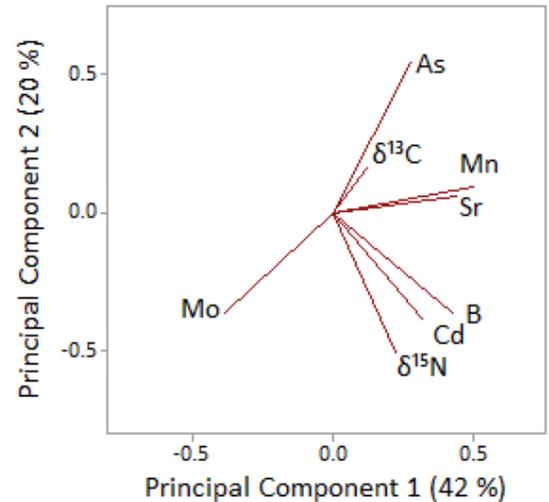
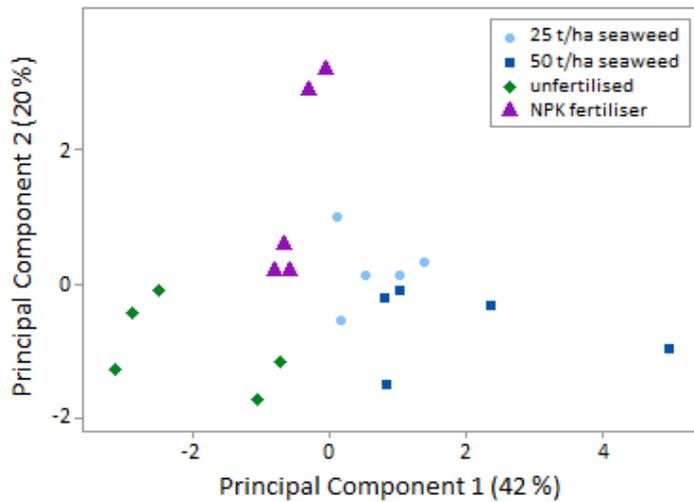
280 A selection of the analytical results of the chemical and isotopic composition of the bere
281 barley is shown in Table 1 (in full in Table S.6, supplementary material). The crop
282 compositions vary subtly from plot to plot. To find which of these differences are
283 characteristic for specific fertilisation treatments and thus important to consider further,
284 principal component analysis (PCA) was performed, revealing systematic differences in the
285 chemical and isotopic composition of grain, husks and straw. In a score plot of principal
286 components 1 and 2 incorporating elemental concentration and isotopic composition
287 results from all measured samples, the samples grouped primarily according to plant part,

288 irrespective of fertilisation treatment, with the closest grouping observed for grain, and a
289 wider spread for straw (Fig. S.1, supplementary material). When performing three separate
290 PCAs, one for each studied plant part (grain, husk and straw), clear grouping based on
291 fertilisation treatment was observable, and $\delta^{15}\text{N}$ and concentrations of B, Mn, As, Sr, Mo,
292 and Cd were identifiable as important parameters for differentiating between treatments
293 (Fig. 1).

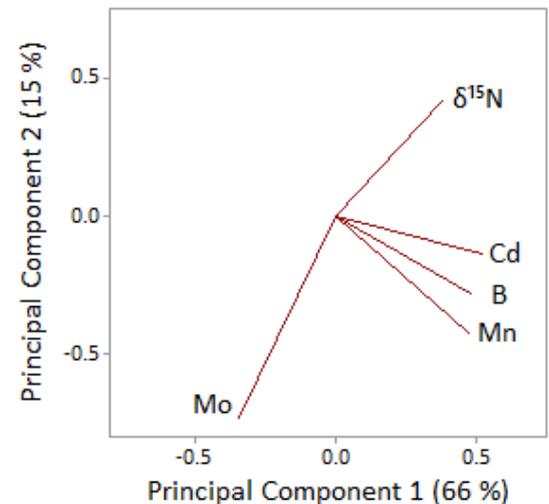
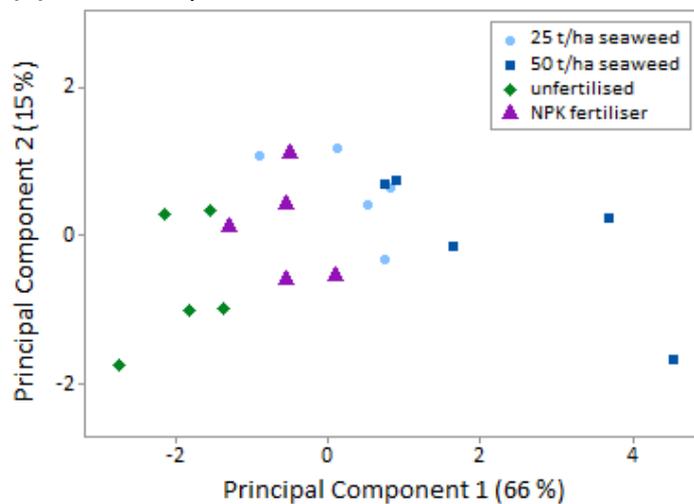
294 The composted seaweed fertiliser had $\delta^{13}\text{C}$ values of $-19.5 \pm 0.2 \text{ ‰}$ (mean $\pm 1\sigma$ of triplicate
295 measurements) and $\delta^{15}\text{N}$ values of $6.7 \pm 0.3 \text{ ‰}$. The results of the analysis of the fertilisers
296 are shown in full in the supplementary material (Table S.7, supplementary material).

297

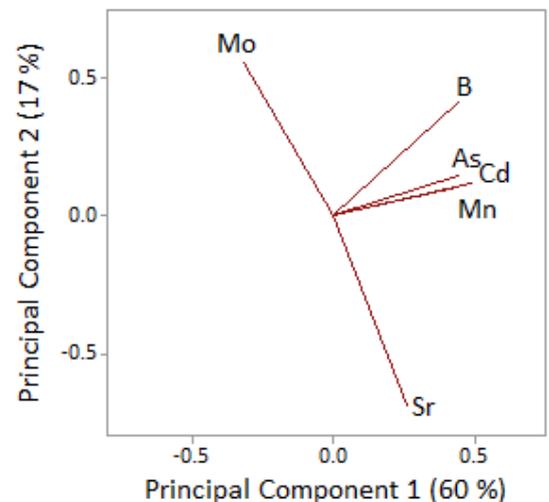
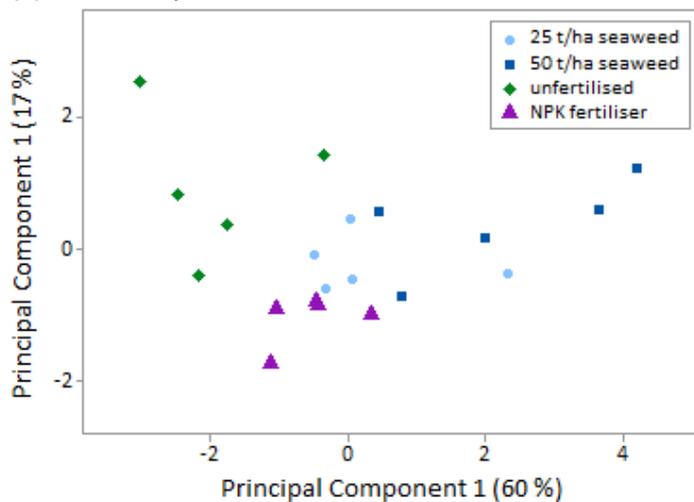
298 (a) grain samples



299 (b) straw samples
300



301 (c) husk samples
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305 **Figure 1** Score plots (left) and loading plots (right) of three principal component analyses of
306 selected element concentrations and isotope ratios (as indicated in the loading plot) for
307 grain, straw and husk samples, indicating the changes induced by the different fertilisation
308 treatments

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Table 1 Selected measured compositional data for seaweed-fertilised, NPK fertilised and unfertilised bere barley grain, husk and straw; values given as weighted averages of seven single measurements (one measurement each for four replicate plots, and triplicate measurements for one replicate plot) for each treatment type $\pm 1\sigma$; letters indicate the results of one-way ANOVA and Tukey post-hoc tests, whereby different letters indicate significant differences ($p < 0.05$) between treatments for each sample type (separately for grain, husk and straw); where no significant differences were found between treatments for the same plant part, no letters are given; in the case of the $\delta^{15}\text{N}$ values for husks where fewer data points were available and no ANOVA was performed, indicated by x; complete set of data reported in Table S.1 in the supplementary material

Sample type	Fertilisation treatment	Mn ($\mu\text{g/g}$)	B ($\mu\text{g/g}$)	As (ng/g)	Sr (ng/g)	Mo (ng/g)	Cd (ng/g)	$\delta^{13}\text{C}$ (‰)	C (%)	$\delta^{15}\text{N}$ (‰)	N (%)	C/N (molar)
grain	no fertiliser	12.1 \pm 1.8 b	1.19 \pm 0.11 bc	13.3 \pm 3.2 b	3.3 \pm 0.2 b	604 \pm 97 a	50.4 \pm 18.7	-27.2 \pm 0.3	40.7 \pm 0.4 a	5.0 \pm 0.4 ab	1.5 \pm 0.1	33 \pm 1
	25 t/ha seaweed	15.6 \pm 1.1 a	1.34 \pm 0.06 ab	25.7 \pm 5.3 ab	4.0 \pm 0.3 a	409 \pm 66 b	48.0 \pm 5.8	-27.1 \pm 0.5	39.7 \pm 0.4 b	5.1 \pm 0.5 a	1.4 \pm 0.1	32 \pm 1
	50 t/ha seaweed	16.4 \pm 2.3 a	1.52 \pm 0.24 a	35.7 \pm 8.6 a	4.2 \pm 0.4 a	413 \pm 57 b	65.9 \pm 11.2	-27.3 \pm 0.5	40.0 \pm 0.8 ab	5.6 \pm 0.3 a	1.6 \pm 0.4	30 \pm 6
	NPK fertiliser	14.6 \pm 0.8 b	1.06 \pm 0.10 c	34.2 \pm 22.0 ab	4.0 \pm 0.3 a	465 \pm 83 ab	46.6 \pm 12.7	-27.1 \pm 0.4	39.8 \pm 0.5 ab	4.3 \pm 0.5 b	1.4 \pm 0.1	34 \pm 4
husk	no fertiliser	14.2 \pm 2.7 b	2.31 \pm 0.36 b	46.8 \pm 22.3 b	8.6 \pm 0.8 b	463 \pm 103 a	39.0 \pm 11.0 b	-27.9 \pm 0.7	44.7 \pm 0.4	3.5 \pm 0.3 x	1.5 \pm 0.2 a	50 \pm 24 b
	25 t/ha seaweed	20.1 \pm 3.4 ab	2.96 \pm 0.70 ab	56.2 \pm 14.4 b	10.3 \pm 1.5 ab	293 \pm 55 b	53.4 \pm 14.3 ab	-27.8 \pm 0.3	44.5 \pm 0.2	3.8 \pm 0.3 x	0.6 \pm 0.5 b	134 \pm 40 a
	50 t/ha seaweed	22.0 \pm 6.9 a	3.98 \pm 0.90 a	100.0 \pm 13.1 a	11.5 \pm 0.8 a	326 \pm 39 b	68.7 \pm 15.7 a	-28.0 \pm 0.4	43.9 \pm 0.5	4.8 \pm 1.3 x	0.8 \pm 0.3 ab	79 \pm 45 ab
	NPK fertiliser	16.7 \pm 2.6 ab	2.10 \pm 0.25 b	55.9 \pm 16.0 b	12.3 \pm 1.3 a	362 \pm 31 ab	46.7 \pm 11.3 ab	-28.0 \pm 0.6	43.9 \pm 0.5	3.0 \pm 0.2 x	1.1 \pm 0.4 ab	52 \pm 22 b
straw	no fertiliser	9.9 \pm 2.9	2.95 \pm 0.34 b	90.8 \pm 16.7	29.0 \pm 4.2	775 \pm 242 a	76.2 \pm 11.0 b	-29.5 \pm 0.2	42.8 \pm 0.5	4.4 \pm 0.2 b	0.3 \pm 0.1	145 \pm 8
	25 t/ha seaweed	16.1 \pm 6.5	3.56 \pm 0.22 b	78.2 \pm 7.1	26.7 \pm 1.7	386 \pm 75 b	101.5 \pm 18.0 b	-29.1 \pm 0.5	42.5 \pm 1.7	5.0 \pm 0.2 a	0.3 \pm 0.0	146 \pm 10
	50 t/ha seaweed	22.7 \pm 12.4	4.64 \pm 1.05 a	98.2 \pm 19.7	28.9 \pm 3.5	404 \pm 83 b	147.8 \pm 27.1 a	-29.4 \pm 0.6	42.3 \pm 1.0	5.5 \pm 0.3 a	0.4 \pm 0.1	138 \pm 14
	NPK fertiliser	14.2 \pm 4.8	3.31 \pm 0.31 b	78.2 \pm 9.8	29.1 \pm 3.8	447 \pm 102 b	99.2 \pm 16.6 b	-29.4 \pm 0.4	42.0 \pm 0.7	4.4 \pm 0.5 b	0.3 \pm 0.0	144 \pm 11

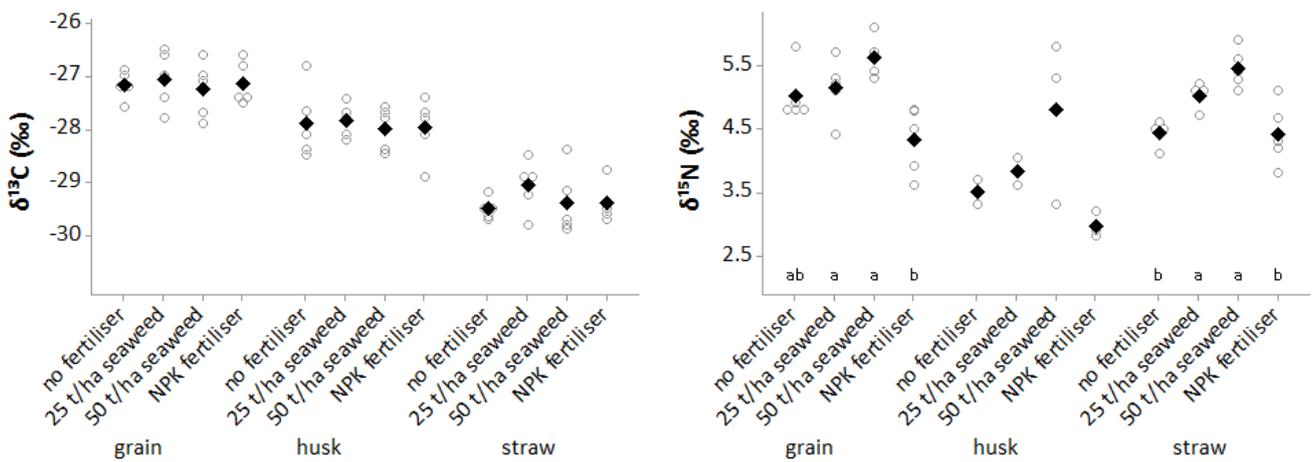
317

317 4.1 Nitrogen ($\delta^{15}\text{N}$) and carbon ($\delta^{13}\text{C}$) stable isotope ratio results

318 The results of the $\delta^{15}\text{N}$ analyses are shown in Table 1 and Fig. 2. Measured $\delta^{15}\text{N}$ values for the 50
 319 t/ha seaweed fertilised barley were significantly elevated when compared to those of the
 320 unfertilised control plots by $0.6 \pm 0.5 \text{ ‰}$ (average $\pm 1\sigma$) in the case of grain (t-test, $p = 0.04$), and by
 321 $1.1 \pm 0.4 \text{ ‰}$ in the case of straw (t-test, $p = 0.001$). Values for 25 t/ha seaweed fertilised barley were
 322 between those of the unfertilised and 50 t/ha seaweed treatment, while the lowest values were for
 323 the NPK treated barley. In husks, highly variable nitrogen concentrations (0.2–1.6 % N; see also
 324 chaff in Bogaard et al., 2007) caused some inaccuracy for husk $\delta^{15}\text{N}$ measurements for which
 325 reason these husk $\delta^{15}\text{N}$ results were excluded here, but are shown in supplementary material
 326 (Table S.6).

327 No significant differences in $\delta^{13}\text{C}$ values were observed between treatments (see Fig. 2), but
 328 significant differences between plant parts were observable, with average grain $\delta^{13}\text{C}$ values
 329 elevated by $0.8 \pm 0.6 \text{ ‰}$ and $2.2 \pm 0.6 \text{ ‰}$ compared to husks and straw, respectively (one-way
 330 ANOVA followed by Tukey indicate the 3 means to be significantly different).

331



332

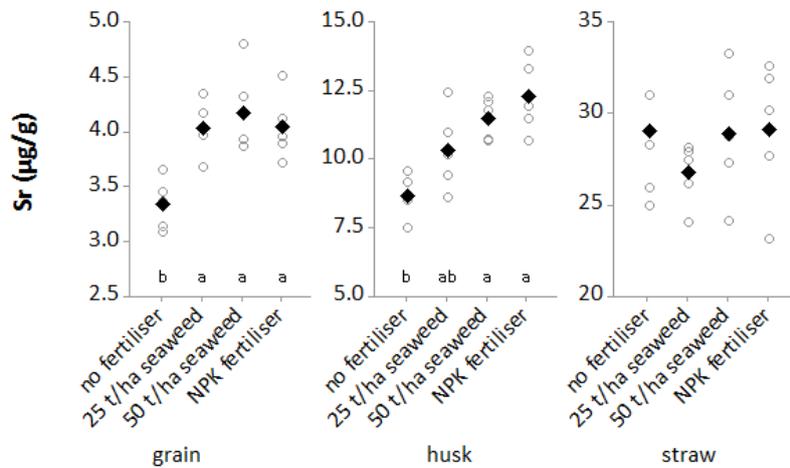
333

334 **Figure 2** Carbon and nitrogen stable isotope ratios ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) in bere barley following various
 335 fertilisation treatments; the circles in each column represent results from five samples (one from
 336 each replicate plot) and the black diamonds indicate the average of these values; within each
 337 column different letters for samples from the same plant part (grain, husk, or straw) indicate
 338 significant differences ($p < 0.05$: one-way ANOVA and Tukey post-hoc tests); in the case of the $\delta^{15}\text{N}$
 339 values for husks fewer data points were available and no ANOVA was performed

340 4.2 Strontium (Sr) concentrations

341 The results of the Sr analyses are given in Table 1 and Fig. 3. Sr concentrations in grain and husks
 342 from 25 t/ha, 50 t/ha seaweed and NPK fertilised plots were elevated by factors of 1.2 to 1.4 (on
 343 average) when compared to grain husks from unfertilised plots (significantly different at $p < 0.05$). In
 344 the case of straw, no significant difference in Sr concentrations was observed between treatment
 345 groups (one-way ANOVA: $F(3,16) = 0.56$, $p = 0.7$). When comparing between plant parts, the highest

346 Sr concentrations were observable in straw (23 to 35 $\mu\text{g/g}$ across all treatments) and the lowest in
 347 unfertilised grain (3.0 to 3.6 $\mu\text{g/g}$).



348

349 **Figure 3** Strontium concentrations in bere barley following various fertilisation treatments; the
 350 circles in each column represent results from five samples (one from each replicate plot) and the
 351 black diamonds indicate the average of these values; within each column different letters for
 352 samples from the same plant part (grain, husk, or straw) indicate significant differences ($p < 0.05$:
 353 one-way ANOVA and Tukey post-hoc tests)

354 4.3 Effect of seaweed fertilisation on other element concentrations

355 Other elements with significantly elevated concentrations in samples from the 50 t/ha seaweed
 356 fertilised plots compared to samples from unfertilised plots included arsenic (As; t-test, $p \leq 0.004$
 357 for husks and grains, but $p = 0.5$ for straw), boron (B; t-test, $p \leq 0.04$ for husk, grain and straw),
 358 manganese (Mn; t-test, $p = 0.02$ for grain, but $p \geq 0.07$ for husk and straw) and cadmium (Cd; t-test,
 359 $p \leq 0.01$ for husk and straw, but $p = 0.2$ for grain).

360 However, in the case of molybdenum (Mo), the opposite was found, whereby concentrations in
 361 unfertilised grain, husk and straw were significantly elevated when compared to their 25 and 50
 362 t/ha seaweed-fertilised and NPK-fertilised counterparts (t-tests, $p \leq 0.04$ for husk, grain and straw;
 363 except husk from NPK plots, where $p = 0.07$). No significant differences in Fe, Cr, Co, Zn or Pb
 364 concentrations were found between 50 t/ha seaweed-fertilised plots and unfertilised plots in grain,
 365 husk and straw (t-tests, all $p \geq 0.1$).

366 5 Discussion

367 5.1 Effect of seaweed fertilisation on plant nitrogen (N)

368 The increases of $0.6 \pm 0.5 \text{‰}$ (in grain) and $1.1 \pm 0.4 \text{‰}$ (in straw) in $\delta^{15}\text{N}$ values may not appear to
 369 be particularly large when compared to the size of a typical trophic level enrichment (i.e. 3 to 5 ‰
 370 in bone collagen; Bocherens and Drucker, 2003; Hedges and Reynard, 2007). However, since this
 371 study was undertaken on soil that had been fertilised in previous years (i.e. already improved soil
 372 with comparatively good initial nutrient status), it is likely that had no previous fertilisation taken
 373 place, or in particularly poor soils, seaweed-fertilisation would have had a greater effect.

374 Additionally, the recovery of intact (though weathered) pieces of seaweed from the trial plots after
375 harvest also indicate long-term effects due to seaweed fertilisation, as further seaweed decay was
376 yet to take place (beyond the end of the trial period). Moreover, compared to historical seaweed-
377 fertilisation practices with rates as high as 50 t/ac (124 t/ha) and multiple applications per year
378 (Fenton, 1997; Russell, 1910; Sauvageau, 1920), the single application fertilisation rates of 25 t/ha
379 and 50 t/ha employed here are still very low. However, the difference between the 25 t/ha and 50
380 t/ha seaweed fertilised plots in this trial indicates that higher seaweed application rates lead to a
381 higher degree of enrichment of ^{15}N . Thus, higher application rates and the repeated application of
382 seaweed within the same season of growth over decades of farming can be expected to lead to
383 higher $\delta^{15}\text{N}$ values.

384 The ^{15}N enrichment observed here appears to be only slightly smaller than that arising from the
385 application of farm-yard manure in comparable short-term experiments (Choi et al., 2006; Fraser et
386 al., 2011), while long-term experiments (over 100 years) with animal manure have led to higher
387 degrees of enrichment (e.g. 9 ‰ in one particular trial; Fraser et al., 2011), giving further indication
388 that effects of long-term fertilisation with seaweed may be similarly substantial. Thus, studies of
389 $\delta^{15}\text{N}$ in archaeological charred cereal grains undertaken to identify past agricultural practices and
390 growing conditions, such as fertilisation with animal manure (e.g. Gron et al., 2017; Kanstrup et al.,
391 2011), should also consider the possibility of seaweed fertilisation, particularly in coastal areas.

392 In order to apply this to the study of consumer skeletal material, it needs to be considered that
393 consumer collagen $\delta^{15}\text{N}$ values are primarily affected by dietary protein $\delta^{15}\text{N}$ values. Here, only
394 total (non-compound-specific) $\delta^{15}\text{N}$ values were determined but it has been shown that
395 fertilisation-induced changes to total $\delta^{15}\text{N}$ reflect changes to the protein $\delta^{15}\text{N}$ composition (Bol et
396 al., 2004; Egle et al., 2008; Styring et al., 2014a, 2014b) .

397 Substantial consumption of seaweed-fertilised crops particularly by weaned herbivores (where the
398 predominant sources of dietary protein are plants; Hedges and Reynard, 2007) but also by
399 omnivores consuming low amounts of protein-rich foods may therefore be assumed to elevate
400 skeletal $\delta^{15}\text{N}$ values compared to consumers of non-fertilised crops (grown under otherwise
401 identical conditions). Even when seaweed-fertilised crops are not directly consumed, elevated $\delta^{15}\text{N}$
402 values of these primary consumers can also be transferred up the food chain (Hedges and Reynard,
403 2007), introducing issues of equifinality both in simple and complex diets. Seaweed-fertilisation
404 may thus cause overestimations of trophic levels throughout the food chain, which may involve
405 both overestimation of the amount of animal products consumed, and overestimation of the
406 trophic level of the consumed animals.

407 5.2 Effect of seaweed fertilisation on plant carbon (C)

408 Since fertilisation with seaweed also introduces marine carbon, this may be expected to have a
409 similar effect on $\delta^{13}\text{C}$ as sea spray, which has been asserted to lead to elevated $\delta^{13}\text{C}$ values in plants
410 because plant roots also take up CO_2 and HCO_3^- from the soil (Göhring et al., 2018). However, no
411 significant differences in $\delta^{13}\text{C}$ values attributable to fertiliser application were observed here. This
412 may be due to the short length of the field trial, but considering the relatively low amount of
413 carbon taken up by plant roots and translocated to the upper parts of the plant compared to that

414 taken up from the atmosphere (Biscoe et al., 1975; Farrar and Jones, 2008; Zamanian et al., 2017), a
415 more significant factor for both seaweed fertilisation and sea spray effects may be salt-stress.
416 Salt stress has been shown to cause elevated $\delta^{13}\text{C}$ values in plants by inducing partial closing of
417 stomata (van Groenigen and van Kessel, 2002), thus introducing what might be interpreted as a
418 more marine isotope ratio without introducing marine carbon. This difference in origin of carbon in
419 plants is of particular relevance to radiocarbon dating due to the marine reservoir effect. However,
420 as no significant differences in $\delta^{13}\text{C}$ due to fertilisation treatments were observed here, these long-
421 term effects are likely comparatively small, and e.g. the systematic differences between $\delta^{13}\text{C}$ values
422 in different plant parts (also previously reported by e.g. Bogaard et al., 2007; Bol et al., 2005;
423 Kanstrup et al., 2011; Sembayran et al., 2008; Serret et al., 2008; Zhao et al., 2001) have a much
424 more immediate relevance for archaeological interpretations.

425 5.3 Effect of seaweed fertilisation on strontium (Sr)

426 Fertilisation with seaweed led to elevated Sr concentrations and Sr/Ca ratios in the fertilised crops
427 (grain and husk). Since the extent to which Sr may substitute for Ca in skeletal bioapatite is affected
428 (at least in part) by dietary Sr concentrations (Bentley, 2006; Sponheimer et al., 2005), these results
429 support suggestions that the elevated Sr concentrations found in some archaeological skeletal
430 material from coastal areas may be due to seaweed fertilisation (Evans et al., 2012; Montgomery et
431 al., 2007, 2003; Montgomery and Evans, 2006).

432 Additionally, the elevated Sr concentrations in grain and husk samples from seaweed-fertilised
433 support hypotheses that strontium isotope ratio $^{87}\text{Sr}/^{86}\text{Sr}$ of crops would become more marine due
434 to seaweed fertilisation (Evans et al., 2012; Montgomery et al., 2007, 2003; Montgomery and
435 Evans, 2006) when growing crops on soils with non-marine Sr isotope ratios. Strontium isotope
436 ratio measurements were not performed for this study, as $^{87}\text{Sr}/^{86}\text{Sr}$ isotope ratios of both seaweed
437 and soil would be expected to be marine due to the close proximity of the trial site to the ocean
438 (Evans et al., 2010; Whipkey et al., 2000). Under these circumstances, no significant differences in
439 $^{87}\text{Sr}/^{86}\text{Sr}$ between seaweed-fertilised and unfertilised crops would be expected.

440 5.4 Effect of seaweed fertilisation on other elements

441 Cd, B, Mn and As concentrations were also elevated in at least some parts (grain, husk or straw) of
442 the seaweed-fertilised barley. It has been reported that As is elevated in soil following seaweed
443 fertilisation, but washes out in subsequent years (Castlehouse et al., 2003), which is consistent with
444 the results presented here. Elevated elemental concentrations due to fertilisation with seaweed
445 appear to be intuitive; however, it should be noted that in several cases, no increase in
446 concentrations were observed (e.g. in the cases of Fe and Pb), while in the case of Mo, lower
447 concentrations were found in seaweed-fertilised crops than in unfertilised crops. Such differences
448 in uptake and translocation are in part related to complex interactions within the soil that affect the
449 solubility and therefore plant uptake of these elements.

450 Particularly the lower concentrations of Mo in fertilised crops (regardless of the type of fertiliser)
451 compared to unfertilised crops in this trial may seem counter-intuitive: Both fertilisers introduce
452 additional Mo to the soil (see Table S.7, supplementary material), and previous studies have shown

453 that when adding only Mo to a Mo deficient soil, an increase in grass Mo concentrations is
454 observable (Johnson et al., 1952). However, in the case of seaweed fertilisation, not only Mo is
455 added to the soil, but a range of elements in various chemical forms that may interact with, and
456 even counteract each other. For example, elevated sulphate concentrations and lower soluble
457 phosphate concentrations may both suppress molybdate uptake, while soils with poor drainage and
458 rich in organic matter generally accumulate soluble Mo (reviewed in Kaiser et al., 2005). The case of
459 Mo therefore serves to illustrate the complexities involved in soil chemistry, element bioavailability
460 and plant uptake mechanisms that can all lead to higher/lower translocation and concentrations in
461 plants. This shows the necessity of experimentally testing assumptions as to how crops are affected
462 by different fertilisers in field trials such as this one, and of considering each element individually.
463 Further study of the effects of seaweed-fertilisation on the trace elemental composition of crops
464 may be of benefit to the development of trace elemental composition analysis of enamel as a
465 means of improving the identification of direct seaweed consumption in complex diets.

466 5.5 Implications for archaeological studies

467 Historical evidence indicates the widespread use of seaweed as a fertiliser across coastal Europe
468 during recent centuries, causing yield increases of comparable extent to fertilisation with animal
469 manure (Hendrick, 1898). As the availability of both animal manure and draught animals have been
470 proposed to be key limiting factors for fertilisation practices in Neolithic Europe (Bogaard, 2012;
471 Gron et al., 2017), it seems plausible (or even likely) that fertilisation with seaweed, which was
472 widely available along the coastline, was practiced from the Neolithic onwards (Bell, 1981; Milner et
473 al., 2004; Schulting et al., 2010). Therefore, the chemical study of skeletal remains needs to
474 consider the effects of fertilisation with seaweed.

475 Previous work has already explored the implications of fertilisation with animal manure for dietary
476 reconstructions with respect to $\delta^{15}\text{N}$ values (e.g. Bogaard et al., 2013, 2007; Styring et al., 2015;
477 Szpak, 2014), and these considerations also apply to seaweed fertilisation, in that consumer trophic
478 levels may be overestimated when fertilisation with seaweed is not accounted for. The direct study
479 of $\delta^{15}\text{N}$ in archaeological charred cereal grains as well as animal remains could be instrumental in
480 resolving problems of equifinality in mixed diets.

481 However, while determining $\delta^{15}\text{N}$ values in archaeological crop samples would aid in dietary
482 reconstructions of animal and human diets, their use in identifying past fertilisation practices is
483 complicated as elevated plant $\delta^{15}\text{N}$ values can arise from a variety of causes (reviewed in Craine et
484 al., 2015). While this study shows that it is likely possible to distinguish between crops fertilised
485 with animal manure and seaweed on the basis of trace element concentrations in modern field
486 trials on the same soil, diagenesis would presumably prevent this from succeeding with
487 archaeological crop samples in most cases.

488 The lack of a significant effect of seaweed fertilisation on crop $\delta^{13}\text{C}$ indicates that short-term
489 fertilisation with seaweed (and/or other marine materials) is unlikely to induce significantly higher
490 $\delta^{13}\text{C}$ values in crops. Hence, e.g. the elevated $\delta^{13}\text{C}$ values found in sheep as compared to cattle in
491 Orkney (Scotland) during the Neolithic and Bronze Age (as discussed in Jones and Mulville, 2016)
492 are perhaps more likely to have arisen from the occasional direct consumption of seaweed (Balasse

493 et al., 2009, 2005; Hansen et al., 2003) rather than from the consumption of marine-fertilised
494 terrestrial plants. Growing fertilised crops requires significantly more labour than the direct
495 consumption of seaweed in coastal areas, and particularly in times of scarcity, animals would have
496 been unlikely to feed primarily on fertilised crops when such crops could instead be consumed by
497 humans. It is therefore important to separate the direct consumption of seaweed (on the one hand)
498 and seaweed-fertilised terrestrial crops (on the other). This may be done by studying $\delta^{13}\text{C}$ values of
499 skeletal material; but when seaweed is only a small part of the total diet, its contribution may well
500 be unidentifiable by $\delta^{13}\text{C}$ alone, and the additional study of trace element concentrations may aid
501 interpretations.

502 This study has also shown that fertilisation with seaweed introduces significant amounts of Sr into
503 the terrestrial food web, which may help explain the elevated Sr concentrations with marine
504 $^{87}\text{Sr}/^{86}\text{Sr}$ ratios observed in some coastal populations (cf. Evans et al., 2012). The elevated Sr/Ca
505 ratios in grain and husks suggest that the Sr/Ca ratio in skeletal material, which has been used as a
506 biochemical indicator of past diet (Peek and Clementz, 2012; Sponheimer et al., 2005; Sponheimer
507 and Lee-Thorp, 2006), is likely also affected by the consumption of seaweed-fertilised crops.
508 Similarly, seaweed-fertilisation of terrestrial crops may complicate attempts to utilise trace element
509 concentrations in tooth enamel to identify seaweed consumption.

510 6 Conclusion

511 This study demonstrates that fertilising terrestrial crops with seaweed can lead to significant
512 changes in plant chemical and isotopic composition, even when fertilisation was only undertaken
513 once, particularly with respect to $\delta^{15}\text{N}$ and Sr concentrations. In the case of $\delta^{15}\text{N}$, an elevation by
514 $0.6 \pm 0.5 \text{ ‰}$ (average $\pm 1\sigma$) in grain and by $1.1 \pm 0.4 \text{ ‰}$ in straw was observed upon fertilisation with
515 50 t/ha seaweed, which is not a substantial increase in trophic level terms, but this likely stacks up
516 over several fertilisation cycles. This effect could then lead to an overestimation of the trophic level
517 of the consumers and their predators in dietary studies. No increase in $\delta^{13}\text{C}$ upon seaweed
518 fertilisation was observed here, indicating that seaweed fertilisation is unlikely to significantly
519 influence $\delta^{13}\text{C}$ values in the skeletal tissues of animal and human consumers.

520 Seaweed fertilisation also led to increased Sr concentrations in barley grain and husk, indicating
521 that seaweed-fertilisation may contribute to long-term enrichment of soil Sr concentrations. This
522 implies that on soils with originally non-marine $^{87}\text{Sr}/^{86}\text{Sr}$ ratios, seaweed-fertilisation may induce
523 more marine Sr isotope ratios in cereal grain. In contrast, depleted concentrations of Mo in
524 seaweed-fertilised barley (when compared to unfertilised barley) indicate that the addition of
525 certain elements to the soil does not necessarily lead to increased translocation into crops. This
526 underlines the importance of testing assumptions and systematically mapping out baseline data
527 using modern field trials to enable accurate archaeological conclusions. Further research into the
528 longer-term effects of seaweed fertilisation on crops has the potential to contribute significantly to
529 our understanding of past coastal populations and their dietary practices.

530 7 Author contributions

- 531 Study conception and literature review: MB
532 Field trial design and implementation: PM, MB, BD, JW, IM
533 Yield evaluation and sample preparation: BD, MB
534 MP-AES, ICP-MS and IRMS measurements: MB, AR, KS
535 PCA, figure preparation and first draft: MB
536 Revision of manuscript: all authors
537 All authors read and approved the final draft prior to submission.

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549 9 Supplementary Material

550 Additional information on the field trial as well as Fig. S.1, Table S.1, Table S.2, Table S.3, Table S.4,
551 Table S.5, Table S.6 and Table S.7 can be found in the online supplementary material to this article
552 at [hyperlink here](#).

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